

1 **Title: Carbon bias from grassy tree misclassification: revealing forest structural**
2 **heterogeneity by integrating GEDI and Sentinel-2**

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10 **Abstract**

11 Tropical forests store roughly half of terrestrial carbon, yet carbon estimates in regenerating and
12 disturbed forests remain highly uncertain. A major source of bias is the prevalence of fast-
13 growing, canopy-forming monocots—such as bamboo, palms, and bananas—that are often
14 misclassified as trees. These “grassy trees” achieve canopy dominance but lack secondary
15 growth, violating woody allometries used in most biomass models. Although NASA’s GEDI
16 mission has transformed large-scale biomass mapping with spaceborne LiDAR, its products rely
17 on coarse plant functional types (PFTs), causing grassy-tree-dominated canopies to be absorbed
18 into evergreen broadleaf tree (EBT) classes. Using a texture-based Sentinel-2 classifier, we
19 isolated bamboo-dominated forests within GEDI EBT products in Xishuangbanna, China. GEDI
20 observations show that bamboo canopies are structurally distinct from tree-dominated forests and
21 lead to systematic carbon overestimation of 20–44 Mg C ha⁻¹ relative to empirical benchmarks.
22 Our framework improves carbon accounting in structurally heterogeneous forests while
23 remaining adaptable for place-based management.

24 **Keywords:** Spaceborne LiDAR, Sentinel-2, Sub-PFT stratification, Carbon Uncertainty, Grassy
25 Trees, Bamboo forests, Forest Monitoring

26 **1. Introduction**

27 Tropical forests are a crucial component of the global carbon cycle, storing approximately 50%
28 of terrestrial carbon. While old-growth forests hold large carbon stocks and biodiversity,
29 regenerating and frequently disturbed forests have expanded from 16.2% to 28.9% of total
30 tropical forest area over the past 30 years, sequestering carbon at rates five times higher than
31 intact forests (Pan et al. 2024). However, their carbon dynamics remain highly uncertain ($\pm 20\%$)
32 despite their increasing importance in land carbon sinks. This uncertainty is amplified in
33 regrowing forests because disturbance and recovery generate fine-scale structural and
34 compositional heterogeneity that challenges mapping and biomass inference.

35 A prominent contributor to this heterogeneity is the prevalence of canopy-forming
36 monocots—large monocots such as bamboo, palms, and bananas—whose canopy architecture
37 and biomass allocation systematically depart from the tree-dominated canopies that plant
38 functional type (PFT) labels like “evergreen broadleaf trees” implicitly represent. These “grassy
39 trees” have reached tree-like canopy dominance but are constrained by the lack of secondary
40 growth like grasses with densely packed herbaceous tissues, often have hollow stems and lower
41 aboveground biomass density than typical trees (Avalos et al. 2022; Zheng and Lu 2025). They
42 can dominate disturbed tropical canopies (e.g., bamboo covering up to 21.3% of southwestern
43 Amazonian forests; (de Carvalho et al. 2013) and may sequester carbon at rates several times
44 higher than typical regrowth forests (Heinrich et al. 2021; Nath et al. 2015). Differences in
45 above- and belowground biomass of grassy trees (Yuen et al. 2017; Navarro et al. 2008;
46 Kotowska 2015; Song et al. 2017) further complicate tree-based biomass estimations where they
47 are abundant.

48 Recent advancements in high-resolution remote sensing offer a path forward in
49 addressing these challenges arising from complex canopy structures associated with grassy tree
50 abundance in disturbed and fragmented landscapes. Sentinel-2 from the European Space
51 Agency’s Copernicus programme (10-m; global since 2015; (Drusch et al. 2012)), coupled with
52 texture-based analysis techniques, provides new opportunities to identify forest areas dominated
53 by bamboo and palms. Unlike traditional spectral classification, texture-based approaches such
54 as the Gray-Level Co-occurrence Matrix (GLCM) capture canopy structural patterns that
55 distinguish grassy tree canopies from those formed by broadleaf woody trees (Humeau-Heurtier
56 2019). For example, sub-meter-resolution Maxar imagery shows that grassy trees such as
57 bamboo form elongated or patchy canopy patterns, in contrast to the rounded crowns of
58 broadleaf woody trees (see Fig.1 and Fig.A.1).

59 These canopy texture cues provide the horizontal context that complements vertical
60 insights from LiDAR missions such as NASA’s Global Ecosystem Dynamics Investigation
61 (GEDI), which delivers global products for canopy height (L2A), canopy cover and vertical
62 structure (L2B), and aboveground biomass density (L4A) (Dubayah et al. 2022). These structural
63 insights are crucial to address two primary barriers that have prevented the integration of grassy
64 trees into forest biomass assessment in previous research. First, forest inventories have
65 historically overlooked these groups due to ambiguous growth forms, leading to
66 underrepresentation in ground-truth data (Fadrique et al. 2020). Second, remote sensing has
67 struggled to accurately identify these functionally unique species within forest canopies. Grassy
68 trees share spectral properties similar to those of broadleaf trees (Li et al. 2019; Yusof et al.
69 2021), while lower-resolution sensors such as Landsat (30 m) and MODIS (250 m) lack the
70 spatial detail needed to detect them (Qi et al. 2022; Venkatappa et al. 2020; Jia et al. 2020). This

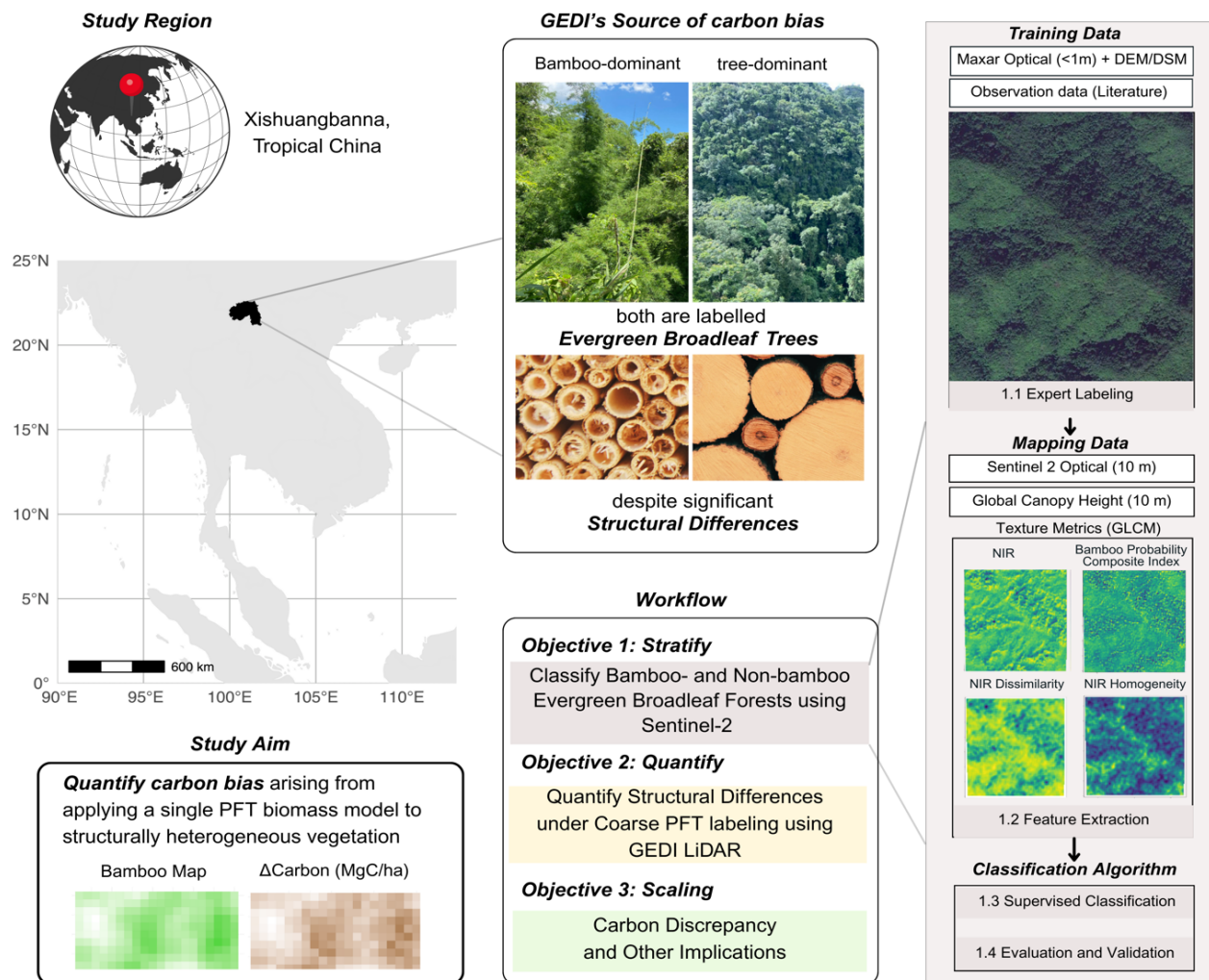
71 spectral and spatial ambiguity has compounded existing ground-based biases, reinforcing
72 uncertainties in forest carbon estimates.

73 However, GEDI LiDAR products have limitations because they stratify global forests
74 into five PFTs—evergreen broadleaf trees (EBT), deciduous broadleaf trees, evergreen
75 needleleaf trees, deciduous needleleaf trees, and a combined grass–shrub–woodland class—
76 informed by MODIS land cover inputs (Kellner et al. 2023). While suitable for global-scale
77 modeling, this coarse classification can obscure critical structural differences among vegetation
78 subtypes, especially in regions with high functional heterogeneity. Applying a single PFT label
79 (e.g., EBT) across diverse canopy structures—such as tree-dominated, palm-rich, or bamboo-
80 dominated forests—can introduce systematic biases in GEDI-based biomass estimates (Bruening
81 et al. 2023; Lin et al. 2025). These issues are particularly acute in tropical Asia and Africa, where
82 calibration data are sparse, and canopy forms often depart from the tree-based assumptions
83 embedded in GEDI biomass models (Dubayah et al. 2022).

84 To address these limitations, we combine GEDI with complementary datasets to improve
85 grassy tree stratification and quantify carbon-estimation bias from mis-labeling grassy trees as
86 EBT. While GEDI provides high-fidelity vertical structure metrics, it lacks the horizontal context
87 needed to disaggregate compositionally diverse forest types. Sentinel-2, despite its limited
88 sensitivity to vertical structure, adds valuable context through its consistent spatial resolution and
89 spectral–textural richness. When paired with texture-based approaches, it can reveal patterns that
90 help distinguish vegetation types otherwise lumped under the same PFT (Humeau-Heurtier 2019;
91 Descals 2023).

92 We conceptualize this integration in Fig.1, where Sentinel-derived compositional context
93 helps subdivide GEDI’s broad PFT class, EBT, into grassy trees (bamboo) and woody trees,

94 identifying areas where structural and biomass deviations are likely and thus require calibration
95 refinement. We then build a specific sub-PFT (e.g., bamboo) classifier for mapping that is
96 designed to generalize. Together, Fig.1 presents the core workflow of this study—combining
97 structure and composition via texture-enhanced satellite imagery and GEDI LiDAR—to stratify
98 heterogeneous forests more accurately and quantify functional misclassification biases in carbon
99 accounting.



100

101 **Fig 1.** Conceptual workflow of this study, combining texture cues from Sentinel-2 with vertical
 102 structure from a global canopy height model (Lang et al. 2022) and GEDI LiDAR metrics to
 103 quantify carbon bias in functionally heterogeneous evergreen broadleaf tree (EBT) forests with
 104 high bamboo abundance in Xishuangbanna.

105 In this study, we focus on bamboo-dominated forests in Xishuangbanna, southwestern
106 China—a region with high bamboo richness, natural abundance, and cultivation, yet currently
107 subsumed under GEDI’s evergreen broadleaf tree (EBT) class. We use this region to demonstrate
108 how combining texture-enhanced Sentinel-2 classification with GEDI structural metrics refines
109 carbon assessments and maps structurally distinct sub-PFT canopies. While this study centers on
110 bamboo, the workflow is designed to be extended to other grassy-tree-dominated systems such as
111 palm forests, as well as other structurally heterogeneous forests, like those with high liana
112 abundance. Specifically, our objectives are outlined in Fig.1:

- 113 1. **Stratify** bamboo-dominated forests from tree-dominated EBT forests and other land-use
114 types using a texture-enhanced Sentinel-2 classification.
- 115 2. **Quantify** structural and biomass differences in GEDI metrics between classified bamboo and
116 non-bamboo EBT forests to evaluate bias from GEDI’s tree-centric EBT biomass model.
- 117 3. **Scale** aboveground biomass density to aboveground carbon density and map the spatial
118 pattern of carbon bias in forests containing sub-PFT bamboo canopies.

119 **2. Methods**

120 **2.1 Study Region**

121 Xishuangbanna, on the northern edge of tropical Asia, hosts true tropical rainforests despite its
122 monsoonal climate (Zhu et al. 2006). A biodiversity hotspot with over 5,000 higher plant species
123 in just 19,690 km², the region has experienced extensive land-use change over recent decades (Li
124 et al. 2008). Forest cover has declined from ~60% in the 1950s due to the expansion of rubber
125 plantations, cash crops, and shifting cultivation (Zhang and Cao 1995; Senf et al. 2013)(Zhang
126 and Cao 1995); (Zhang and Cao 1995; Senf et al. 2013), giving rise to widespread secondary

127 vegetation—especially bamboo forests (*Dendrocalamus membranaceus*, *Cephalostachyum*
128 *pergracile*, etc., Yang et al. 2008)—now common along forest edges and within fragmented
129 landscapes (bamboo in Fig.1).

130 Xishuangbanna presents an ideal setting to develop and evaluate our method, given its
131 highly heterogeneous forest matrix, strong bamboo presence, and the cultural significance of
132 bamboo among the local Dai population (Wang et al. 2002). We focus on this region to assess
133 how the broad classification of bamboo as evergreen broadleaf trees in GEDI products can affect
134 biomass estimation and carbon accounting.

135 **2.2 Remote Sensing and Analytical Workflow**

136 To accomplish the three objectives of our study, we designed an integrated remote sensing and
137 analytical workflow that includes training data compilation, texture analysis, classification and
138 accuracy assessment, evaluation of GEDI’s limitations, and the quantification of carbon
139 discrepancies (Fig.1). Below, we describe each component of this workflow in detail.

140 **2.2.1 Datasets and Preprocessing**

141 We aligned GEDI footprints to the 10-m Sentinel-2 grid by assigning each footprint to the 10-m
142 pixel containing its geolocation (footprint center). Land-cover labels were extracted at that pixel
143 and linked to footprint-level GEDI metrics for subsequent analyses.

144 ***Sentinel-2 (Optical, 10 m resolution)***: We obtained Sentinel-2 imagery from Google Earth
145 Engine (COPERNICUS/S2_harmonized) and generated four seasonal composites (January–
146 March, April–June, July–September, and October–December) for the year 2021, selecting scenes
147 with less than 5% cloud cover. Each composite included nine spectral bands (B2–B8, B11, B12)
148 and was processed using QA60 and s2cloudless cloud masks to reduce atmospheric noise.

149 ***The Global Canopy Height Model*** (10 m, Lang et al. 2022): a probabilistic deep learning model
150 that fuses sparse height data from GEDI LiDAR with dense optical satellite images from
151 Sentinel-2.

152 ***GEDI (LiDAR, 25 m resolution)***: GEDI datasets provided canopy height and relative height
153 metrics (Level 2A), canopy cover and vertical structure (Level 2B), and aboveground biomass
154 density (Level 4A). All GEDI footprints (25-m resolution) were aligned to 10-meter Sentinel-2
155 composites for value extraction and analysis.

156 **2.2.2 Training data**

157 We compiled a training dataset encompassing diverse vegetation and land-use categories,
158 including bamboo, tree-dominated evergreen forests, rubber plantations, fruit and tea plantations,
159 farms, construction areas, and water bodies. Labels were manually interpreted from 2021 Google
160 Earth imagery (licensed Maxar submeter imagery), chosen for its minimal cloud cover and
161 alignment with GEDI and Sentinel-2 temporal windows.

162 We reliably identified bamboo based on its distinct texture and canopy structure (Fig.1),
163 supported by expert knowledge of its landscape distribution—such as along riparian corridors,
164 within rubber plantations, and near forest edges—and by regional vegetation accounts (Zhu
165 2006; Zhu et al. 2015) describing *Dendrocalamus membranaceus* forests. We identified other
166 vegetation types using a combination of phenological signals (e.g., seasonal reflectance in rice
167 paddies), spatial context (e.g., farms near villages), auxiliary map labels (e.g., “Xishuangbanna
168 Tropical Botanical Garden”), and literature-based spatial descriptions (Min et al. 2019; Li et al.
169 2008), drawing from multiple lines of evidence.

170 In Appendix A, we summarized the training data’s sample size and geographic coverage
171 in Table A.1 and Fig A.1 and detailed the visual interpretation criteria used for each land-use

172 type in Table A.2. Fig A.2 shows the strong spectral separability of the October-September
173 composite for Xishuangbanna in 2021.

174 **2.2.3 Feature Extraction**

175 For each labeled polygon, we extracted a comprehensive suite of features that capture spectral,
176 textural, and contextual information (Fig. A.3). *Spectral features* include surface reflectance
177 from selected Sentinel-2 bands (e.g., red, red-edge, SWIR), along with vegetation indices that
178 characterize greenness, moisture, and phenology: Normalized Difference Vegetation Index
179 ($NDVI = (NIR - Red)/(NIR + Red)$), and the Bamboo Phenological Characteristic Index (BPCI =
180 $Red\ Edge\ 1 / (NIR - NIR / SWIR\ 1)$) (Huang et al. 2024). These features were computed from the
181 seasonal composite imagery described earlier.

182 **2.2.4 Texture Metrics**

183 We applied Gray-Level Co-occurrence Matrix (GLCM) techniques to characterize the spatial
184 distribution of pixel intensities, capturing distinct texture patterns associated with bamboo's
185 unique leaf morphology and clumping architecture. Metrics such as pixel dissimilarity,
186 homogeneity, and entropy have proven useful in distinguishing different species within mixed
187 canopies (Mohammadpour et al. 2022). To identify the most informative spectral regions, we
188 examined how texture metrics varied across different vegetation types—especially between
189 bamboo-, tree-dominated forests, and rubber plantations. This analysis reduces dimensional
190 redundancy and minimizes overfitting risks by selecting texture features (Band 8 dissimilarity
191 and B8 homogeneity computed using a 5×5-pixel sliding window) that are less correlated yet
192 discriminative.

193 **2.2.5 Classification algorithm and accuracy assessment**

194 We trained and evaluated two machine learning classifiers: Random Forest (RF) and
195 Convolutional Neural Networks (CNN) (detailed in Appendix B). RF is robust to noise and well-
196 suited for integrating heterogeneous feature sets, while CNN enables pixel-based classification
197 with spatial context sensitivity. To compare their performance, we withheld a subset of ground-
198 truth data and applied both classifiers independently.

199 To assess classification accuracy, we conducted visual cross-validation by randomly
200 sampling 100 points within each land-use class and verifying their true identity in Google Earth
201 Pro (v7.3.6, Google Earth 2025) using high-resolution imagery from 2020 to 2022. Confusion
202 matrices summarizing the results of this photo-interpretation-based validation are presented in
203 Table C.1. We report standard accuracy metrics for the better-performing RF classifier: overall
204 accuracy (the proportion of correctly classified cases), class-wise sensitivity (recall; true positive
205 rate) and specificity (true negative rate), and the F1 score (which penalizes both missed
206 detections and false positives).

207 **2.2.6 GEDI overlay**

208 To evaluate the limitations of GEDI-derived products in distinguishing bamboo-dominated
209 forests, we overlaid GEDI Level 2A (canopy height), Level 2B (canopy cover and vertical
210 structure), and Level 4A (aboveground biomass) with our classified land cover map (Figure C.1).
211 For each GEDI footprint intersecting a bamboo-classified pixel, we extracted its associated
212 canopy height, structural, and biomass metrics. For comparison, the same set of variables was
213 extracted from footprints located within tree-dominated forests and other land-use classes.

214 To account for classification uncertainty—particularly confusion between bamboo and
215 tree categories—we treated the validation-derived misclassification rates as an observation
216 model (confusion matrix; Table C.1) and applied a Bayesian measurement-error correction to

217 infer the posterior true area of each land-cover class (detailed in Appendix C). This yields
218 classification-error-adjusted estimates of the total extent of bamboo-dominated forests and other
219 vegetation types across Xishuangbanna.

220 At the footprint level, we propagated this classification uncertainty by assigning each
221 GEDI footprint a posterior probability of belonging to bamboo versus tree conditional on its
222 mapped label, and then reweighted footprint-level summaries accordingly. This probabilistic
223 reweighting enables more robust estimation of biomass distributions for bamboo- and tree-
224 dominated forests and quantifies the extent to which biomass estimates are biased when bamboo-
225 dominated canopies are treated as evergreen broadleaf forests under the default GEDI L4A EBT
226 biomass model.

227 **2.2.7 Aboveground biomass bias from GEDI EBT model**

228 GEDI Level 4A aboveground biomass density (AGBD, Mg ha⁻¹) estimates in Xishuangbanna are
229 derived from a global EBT biomass model parameterized on GEDI relative height (RH) metrics
230 (Kellner et al. 2023):

$$231 \quad AGBD = 1.113 \times \left(-104.965 + 6.802\sqrt{RH50 + 100} + 3.995 \times \sqrt{RH98 + 100} \right)^2 \quad (Eq. 1)$$

232 RH metrics summarize the vertical structure of a LiDAR footprint as percentiles of the
233 cumulative GEDI waveform energy returned from height z . RH98 approximates the upper
234 canopy envelope (near-maximum vegetation height), whereas RH50 represents the median
235 height of returned vegetation material, capturing the vertical distribution of canopy structure.

236 However, this EBT model is applied uniformly across evergreen broadleaf forests,
237 including bamboo-dominated stands that differ fundamentally from trees in growth form and
238 vertical structure. To test whether this unstratified application induces systematic overestimation
239 and to quantify its effect size, we (1) compared GEDI Level 2A/2B structural metrics between

240 bamboo- and tree-dominated areas classified by our random-forest land-cover map, and (2) used
241 empirical bamboo biomass benchmarks from the same ecoregion (Yuen et al. 2017) as external
242 references for bias quantification (see Table D.1). Given strong management and land-use
243 contrasts, we further separated bamboo into forest versus fallow contexts to obtain context-
244 specific benchmarks and bias estimates.

245 **2.3 GEDI structural comparison**

246 We extracted GEDI Level 2A and 2B metrics from footprints whose surrounding area was
247 classified as bamboo or non-bamboo EBT forests (trees). We compared their canopy height
248 (RH98), relative height profile, plant area index (PAI), and plant area volume index (PAVD)
249 throughout the canopy, to reveal structural differences that are masked when all these forests are
250 labelled as a single EBT class in GEDI-based products. These structural contrasts motivate
251 subsequent bias analysis.

252 *Empirical benchmarks for bamboo aboveground biomass and carbon density*

253 Established methods for estimating individual bamboo biomass often are typically based on
254 diameter and height of clumpy tropical bamboos (Xayalath et al. 2019; Camargo García et al.
255 2023; Huy et al. 2019). Unlike trees, bamboo culms do not exhibit continuous secondary growth,
256 and stand-level biomass depends strongly on age structure and culm turnover rather than height
257 alone (Yen 2016). These demographic properties cannot be retrieved from GEDI waveform
258 metrics. Our objective is therefore not to recalibrate bamboo allometry, but to evaluate whether a
259 tree-based RH model produces systematic overestimation on bamboo and to quantify its effect
260 size relative to empirically observed ranges.

261 Accordingly, we used empirical aboveground carbon density estimates synthesized by
262 (Yuen et al. 2017) for tropical bamboos in Indochina (Myanmar, Laos, Thailand, Vietnam, and

263 India), a region biogeographically adjacent to Xishuangbanna and characterized by comparable
264 species composition. Because Xishuangbanna and Indochina share a contiguous Indo-Burma
265 floristic context with overlapping bamboo taxa and broadly similar monsoonal tropical–
266 subtropical climates (Zhu et al. 2006), we expect these synthesized carbon-density benchmarks
267 to be broadly transferable to bamboo-dominated stands in our study region.

268 Reported summary statistics (mean, standard deviation, and max) were used to derive
269 confidence intervals for comparison with GEDI L4A footprint-level estimates (detailed in
270 Appendix D). To reflect agroforestry mosaics abundant with secondary forests in
271 Xishuangbanna, bamboo pixels were classified as forest or fallow based on a 3×3 neighborhood
272 rule: bamboo cells surrounded by ≥ 5 bamboo or tree neighbors were classified as forest
273 bamboo; otherwise as fallow bamboo.

274 For bamboo footprints, we compared GEDI L4A AGBD estimates with empirical
275 bamboo benchmarks stratified by bamboo context $i \in \{forest, fallow\}$, and quantified
276 aboveground carbon bias as:

$$277 \quad \Delta C_i = AGC_{EBT,i} - AGC_{empirical,i} \text{ (Eq. 2)}$$

278 where aboveground carbon density (AGC) was derived by applying a constant carbon fraction
279 (CF = 0.5) to AGBD for both bamboo and trees. Although CF varies among species and tissues,
280 it has been inconsistently treated across the bamboo literature, whereas CF = 0.5 is widely used
281 as a default in forest carbon accounting (Pan et al. 2011). We therefore adopt a single CF to
282 ensure comparability between GEDI-derived AGBD and empirical benchmarks, noting that
283 alternative defaults (e.g., CF = 0.47; IPCC 2006) would linearly rescale ΔC without affecting its
284 sign. Positive ΔC values indicate overestimation of bamboo carbon by GEDI's EBT model
285 relative to empirical benchmarks.

286 **Spatial extrapolation of carbon bias across Xishuangbanna**

287 To characterize the spatial distribution of aboveground carbon bias (ΔC) across bamboo-rich
288 landscapes in Xishuangbanna, footprint-level estimates must be extrapolated to continuous pixel-
289 based surfaces. To ensure internal consistency with GEDI's operational biomass products, this
290 extrapolation was conducted entirely within the framework of the GEDI EBT biomass model
291 (Eq. 1). Because the EBT model requires both RH50 and RH98 as inputs, we first quantified the
292 statistical relationship between RH50 and RH98 separately for bamboo and tree footprints using
293 GEDI Level 2A data. We applied the class-specific RH50–RH98 relationships to the 10-m global
294 canopy height model (Lang et al. 2022) as an operational approximation to enable internally
295 consistent regional mapping within our framework. Because the Lang et al. (2022) product is
296 generated by fusing sparse GEDI Level 2 RH98 observations with dense Sentinel-2 optical data,
297 it provides a GEDI-anchored proxy for canopy stature; nevertheless, it is not a one-to-one
298 substitute for footprint-level RH98. Any residual mismatch between this 10-m canopy height
299 layer and GEDI RH98 over overlapping footprints in Xishuangbanna is therefore treated as a
300 limitation of our extrapolation. The estimated RH50 and RH98 were then used to compute pixel-
301 level AGBD and aboveground carbon density using the EBT biomass model (Eq. 1).

302 The analytical form of the EBT biomass model (Eq. 1) does not capture all calibration
303 steps and ancillary corrections used to generate official GEDI Level 4A biomass estimates. To
304 quantify the systematic error introduced by applying this simplified RH-based formulation alone,
305 we compared AGBD estimates derived from Eq. 1—using RH metrics from GEDI Level 2
306 data—with published Level 4A AGBD values for footprints where Level 2 and Level 4
307 observations overlapped.

308 Finally, we applied Eq. 1 across the entire Xishuangbanna region and converted AGBD
309 to aboveground carbon density. To illustrate the spatial magnitude of overestimation, carbon bias
310 was expressed as a percentage of EBT-estimated carbon density and mapped separately for forest
311 and fallow bamboo areas:

$$312 \quad \%Bias = \frac{\Delta C_i}{AGC_{EBT,i}} \times 100 \text{ (Eq. 3)}$$

313 This multi-step extrapolation introduces three main sources of uncertainty: (1)
314 misclassification errors from the random-forest land-cover map, (2) uncertainty in the class-
315 specific RH50–RH98 regression models, and (3) systematic error arising from applying the
316 simplified EBT biomass formulation (Eq. 1). These uncertainties were explicitly evaluated and
317 propagated in subsequent analyses (see Results).

318 **3. Results**

319 **3.1 Reliable sub-PFT land-cover classification enables robust bamboo detection**

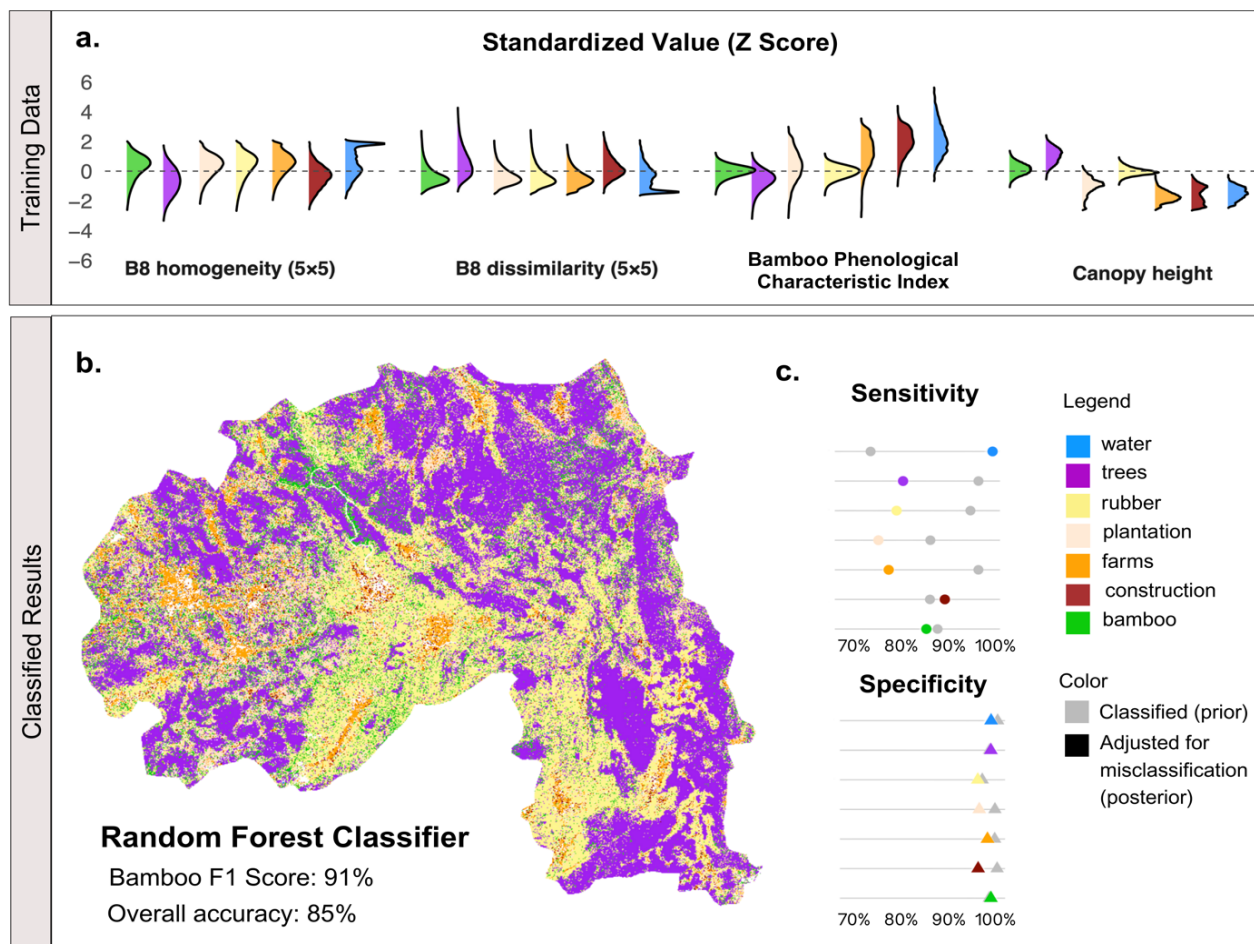
320 The random forest classifier achieved strong overall performance in distinguishing bamboo from
321 other land-cover classes across Xishuangbanna (Fig. 2). Feature distributions derived from
322 spectral texture (Band 8—Near Infrared’s homogeneity and dissimilarity), canopy height, and
323 the Bamboo Phenological Characteristic Index (BPCI) show clear separation among vegetation
324 classes in the training data (Fig. 2a), indicating that these features capture complementary
325 structural and phenological information relevant for bamboo detection.

326 The resulting land-cover map reveals spatially coherent patterns of bamboo, trees, rubber
327 plantations, and agricultural land consistent with known landscape organization in the region
328 (Fig. 2b). For instance, bamboo-dominated riparian belts are apparent along the Lancang River
329 corridor, consistent with reported patterns of river-associated bamboo stands in the region.
330 Beyond riparian zones, bamboo also occurs as patchy elements embedded within rubber

331 plantations and farmland mosaics, consistent with local smallholder management practices.

332 These spatial patterns are further supported by high-resolution aerial maps (Fig A.1).

333 Classification performance was strong despite the highly fragmented tropical landscape,
334 with an overall accuracy of 85% and a bamboo F1 score of 91% (Fig. 2b), indicating reliable
335 bamboo detection with low omission. Nonetheless, confusion among spectrally and structurally
336 similar vegetation classes—most notably bamboo, trees, and rubber plantations—remained non-
337 negligible, consistent with frequent co-occurrence and fine-scale mixing of these vegetation
338 types in agroforestry mosaics. Class-specific sensitivity and specificity (Fig. 2c) further indicate
339 that residual misclassification persists even under a well-performing classifier. We therefore
340 explicitly propagated this classification uncertainty in subsequent analyses using a Bayesian
341 posterior resampling framework and Monte Carlo reassignment of class labels, rather than
342 treating the classified map as error-free. The posterior updating methods and results are detailed
343 in Appendix C.



344 **Fig 2. Sub-PFT mapping reveals hidden vegetation heterogeneity in Xishuangbanna. a)**
345 Standardized distributions (Z-scores) of four key predictors in the training data—Band 8 (near-
346 infrared) homogeneity and dissimilarity (5×5-pixel window), the Bamboo Phenological
347 Characteristic Index, and canopy height—show that the seven mapped classes occupy distinct
348 regions of feature space, indicating that the training data and selected predictors provide strong
349 discriminatory power for subsequent classification. **b)** Using Sentinel-2 spectral indices and
350 GLCM texture features, we classified sub-PFT vegetation and land-cover types across
351 Xishuangbanna with a Random Forest model. The resulting 10-m map reveals fine-scale mosaics
352 of bamboo patches embedded within forest, rubber, and agricultural matrices—heterogeneity
353 typically masked by the EBT forest label in GEDI-based products. **c)** Class-wise sensitivity

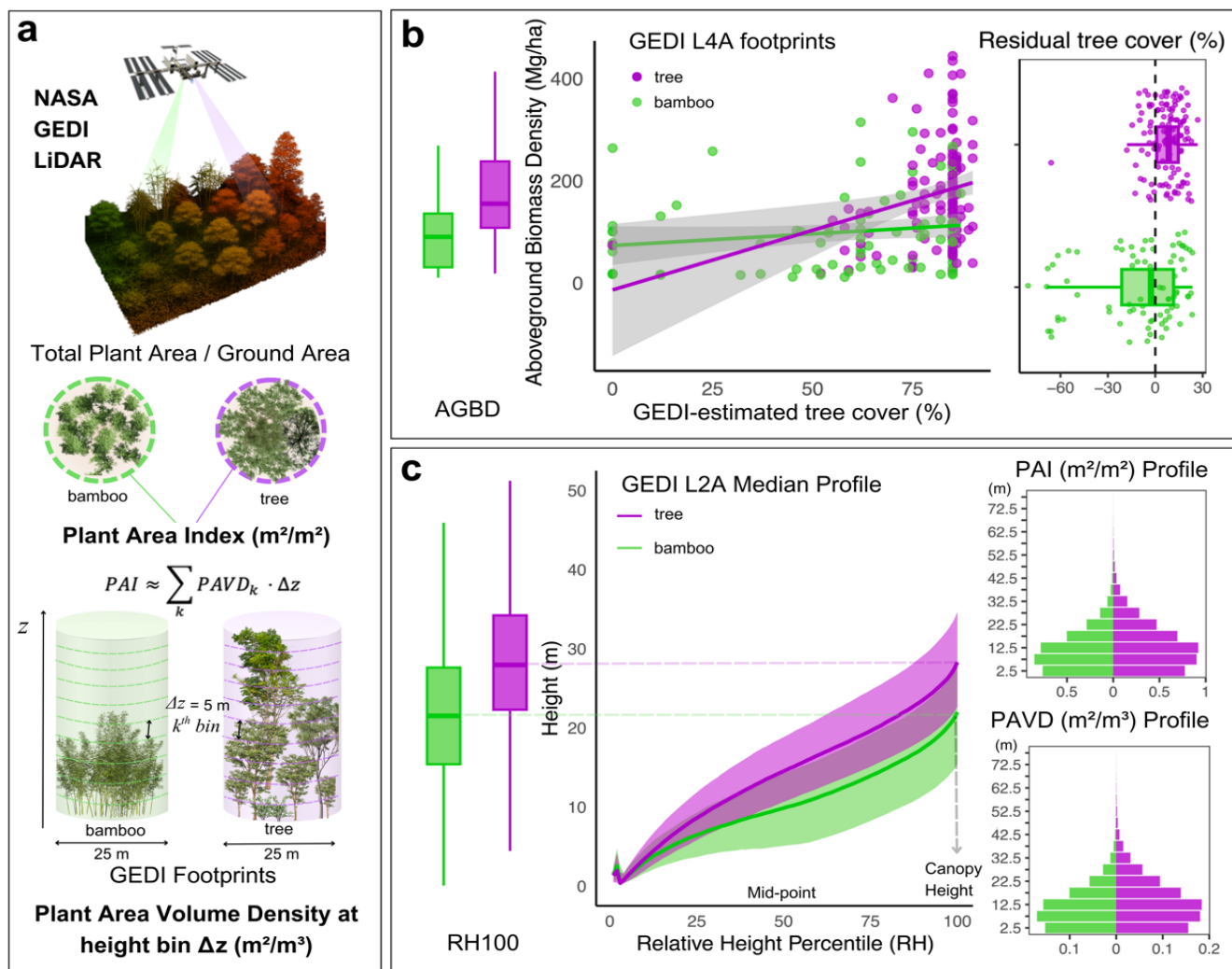
354 (circles) and specificity (triangles) estimated from 70/30 hold-out validation during training
355 (grey symbols) and from visually interpreted sample points drawn from the classified map
356 (colored symbols), which we used to adjust for misclassification. Bamboo achieved an F1 score
357 of 0.91 (overall accuracy 0.85), enabling reliable downstream quantification of bamboo cover
358 and configuration.

359 **3.2 GEDI reveals systematic structural differences between bamboo- and tree-dominated**
360 **forests**

361 Building on the validated land-cover classification and explicitly quantified uncertainty, we
362 overlaid the classification map with GEDI LiDAR footprints to test whether bamboo- and tree-
363 dominated forests exhibit systematic differences in canopy structure and biomass-related metrics
364 as directly observed by GEDI.

365 To aid interpretation of subsequent results, Fig. 3a illustrates how GEDI LiDAR
366 footprint-level observations characterize canopy structure. GEDI footprints sample forest
367 structure within circular columns approximately 25 m in diameter, capturing the vertical
368 distribution of vegetation from the ground to the top of the canopy. Within each footprint, the
369 returned waveform resolves a vertical profile of vegetation material, allowing both canopy height
370 and the allocation of plant material along the vertical axis to be directly observed.

371 As shown in Fig. 3a, the vertical distribution of plant material is summarized by plant
372 area volume density (PAVD) across successive height bins (Δz), while plant area index (PAI)
373 reflects the cumulative plant area per unit ground area within the footprint. Conceptually, PAVD
374 describes how plant material is distributed with height, whereas PAI indicates how plant biomass
375 accumulates vertically, analogous to leaf area index but incorporating both foliage and woody
376 components detected by the LiDAR waveform. Together, these GEDI-derived quantities provide
377 an intuitive framework for interpreting the structural contrasts between bamboo- and tree-
378 dominated forests shown in Fig. 3b–c and Table 1.



379 **Fig 3. Structural differences between bamboo- and tree-dominated forests from GEDI. (a)**

380 Conceptual illustration of how GEDI waveforms capture vertical canopy structure within each

381 ~ 25 m footprint, from which canopy height, plant area index (PAI), and plant area volume

382 density (PAVD) profiles are derived (see Wang et al. 2025). **(b)** Bamboo-dominated footprints

383 have lower AGBD than tree-dominated footprints. Residual tree cover is defined as GEDI-

384 reported tree cover minus the tree-cover value expected for a tree-dominated footprint with the

385 same AGBD; values are consistently negative for bamboo, indicating its cover–biomass

386 decoupling relative to tree-dominated forests. **(c)** Vertical structure differs strongly by vegetation

387 type: bamboo has lower canopy heights (RH98) and a lower upper-canopy profile than trees,

388 with broadly overlapping lower–mid canopy layers. PAI and PAVD profiles are consistent with
 389 these canopy-structure contrasts.

Panel A. Raw GEDI observations (classified bamboo vs trees)					
Metric	n (Bamboo)	Median (Bamboo)	n (Trees)	Median (Trees)	Δ Median (Trees – Bamboo)
Aboveground Biomass Density (Mg ha ⁻¹)	81	90.29	121	154.41	64.12
Canopy Height (m)	31,067	21.87	123,142	28.39	6.52
Plant Area Index (m ² m ⁻²)	43,538	3.28	171,085	4.08	0.81
Total Vegetation Volume (m ³ m ⁻²)	43,508	0.74	171,056	0.90	0.15
Panel B. Misclassification-aware contrasts (posterior Monte Carlo; n = 1000 iterations)					
Metric	Δ Median (MC, 2.5–97.5%)	MC Median	MC $p < 0.05$		
Aboveground Biomass Density (Mg ha ⁻¹)	44.67–66.46	54.37	1.00		
Canopy Height (m)	5.19–5.32	5.26	1.00		
Plant Area Index (m ² m ⁻²)	0.64–0.67	0.65	1.00		
Total Vegetation Volume (m ³ m ⁻²)	0.12–0.13	0.12	1.00		

390 **Table 1. Summary of structural and biomass differences between bamboo- and tree-**
 391 **dominated evergreen broadleaf forests as observed by GEDI LiDAR.** Raw footprint-level
 392 medians (Panel A) are contrasted with misclassification-aware median differences obtained from
 393 posterior Monte Carlo resampling (Panel B; 1,000 iterations). All differences are expressed as
 394 Trees – Bamboo, and the fraction of Monte Carlo iterations with $p < 0.05$ quantifies the
 395 robustness of each contrast to classification uncertainty.

396 Despite being processed using the same EBT biomass model, bamboo- and tree-
397 dominated GEDI footprints exhibited markedly different Level 4A aboveground biomass density
398 patterns (Fig. 3b). We use residual tree cover as a diagnostic of structural inconsistency: for each
399 footprint, we calculate GEDI-reported tree cover minus the tree cover expected at the same
400 AGBD from the unstratified AGBD–tree-cover relationship (i.e., treating all EBT as structurally
401 equivalent). In tree-dominated footprints, AGBD increased strongly with GEDI-estimated tree
402 cover, which was predominantly high (> 50%), consistent with the model’s underlying
403 assumption that canopy cover and vertical extent scale with woody biomass. In contrast,
404 bamboo-dominated footprints showed substantially lower AGBD (median 42% lower) and a
405 weak cover-biomass coupling: similar AGBD values occurred across both low (< 25%) and high
406 (> 50%) tree cover. Accordingly, residual tree cover was predominantly negative for bamboo
407 footprints and positive for tree footprints, indicating a systematic decoupling between cover and
408 biomass in bamboo that arises from GEDI observations alone, prior to any comparison with
409 external biomass benchmarks. Altogether, GEDI tree cover is not an equivalent proxy for
410 biomass in bamboo-dominated EBT forests as it is in tree-dominated EBT forests.

411 GEDI Level 2A metrics further indicate systematic differences in vertical canopy
412 structure between bamboo- and tree-dominated forests (Fig. 3c; Table 1). Bamboo footprints
413 exhibited lower canopy height, with median canopy height ~23% lower than tree footprints.
414 Relative height (RH) profiles show that bamboo canopies concentrate vegetation material at
415 lower heights, resulting in a more compressed vertical distribution, whereas tree canopies exhibit
416 greater vertical extension, and more plant material allocated to upper canopy layers. Consistent
417 with these patterns, bamboo footprints showed lower total PAI and total vegetation volume, and
418 distinct PAVD profiles across height bins. Together, these GEDI-observed structural contrasts

419 provide a mechanistic explanation for the divergent biomass patterns in Fig. 3b and underscore
420 that tree-based structural predictors embedded in the EBT model do not adequately represent
421 bamboo-dominated forests.

422 To assess the sensitivity of these structural contrasts and biomass differences to land-
423 cover misclassification, we propagated classification uncertainty using posterior resampling and
424 Monte Carlo simulations. Across all four GEDI-derived metrics, median differences between
425 tree- and bamboo-dominated footprints remained consistently positive and statistically
426 significant in all Monte Carlo iterations (proportion of iterations with $p < 0.05 = 1.0$; Table 2).
427 Accounting for misclassification reduced the magnitude of the estimated differences relative to
428 naïve estimates—for example, the median tree–bamboo difference in aboveground biomass
429 density decreased from 64.1 Mg ha⁻¹ to 54.4 Mg ha⁻¹—but did not alter their direction or overall
430 effect size. Similar patterns were observed for canopy height, plant area index, and total
431 vegetation volume (Fig C.2). These results indicate that the observed structural differences are
432 robust to plausible classification errors and are not driven by misclassification alone.

433 **3.3 Systematic carbon overestimation in bamboo forests by GEDI EBT models**

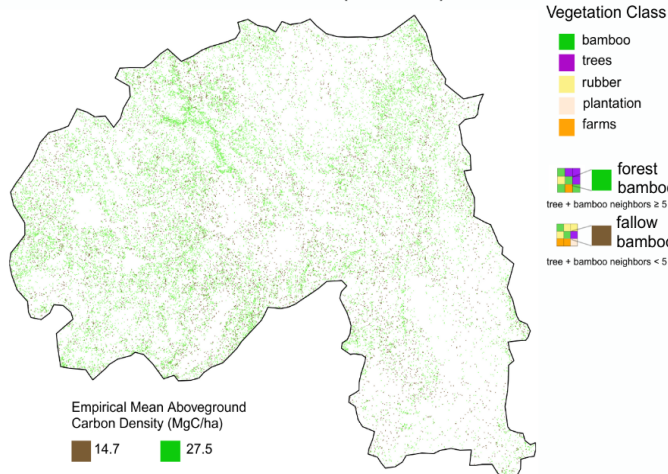
434 Building on the GEDI-observed structural differences between bamboo- and tree-dominated
435 forests (Fig. 3), we next quantified the magnitude of biomass and carbon bias arising from
436 applying the EBT model to bamboo-dominated systems using empirical benchmarks. Bamboo-
437 dominated footprints were stratified into forest and fallow contexts based on local neighborhood
438 composition, reflecting the heterogeneous agroforestry landscape of Xishuangbanna (Fig. 4a).

Panel A. Class-specific RH50–RH98 regressions						
Class	Model	β_0	β_1	n	R ²	RMSE (m)
Trees	$RH50 = \beta_0 + \beta_1 \cdot RH98$	-2.316	0.628	123,142	0.675	3.74
Bamboo	$RH50 = \beta_0 + \beta_1 \cdot RH98$	-3.872	0.649	31,067	0.745	3.35
Panel B. Discrepancy between RH-only EBT re-estimation and published GEDI L4A carbon at overlapping footprints						
Residual (Mg C ha ⁻¹): $\Delta C_{EBT} = C_{est} - C_{L4A}$; $C_{est} = 0.5 \times AGBD_{EBT}(RH50, RH98)$						
Class	n	Mean residual			SD	
Bamboo	81	5.31			51.52	
Trees	121	3.97			44.49	
Panel C. Misclassification-aware bamboo carbon bias relative to empirical benchmarks						
Naive (no-correction) mean bias (Mg C/ha): Bamboo Fallow: 43.7; Bamboo Forest: 19.9						
Bamboo context	Component	Mean bias (Mg C/ha)	Median	2.5%	97.5%	
Fallow bamboo	posterior-resampled bamboo bias (MC)	44.1	43.7	29.6	59.0	
Forest bamboo	posterior-resampled bamboo bias (MC)	19.8	19.8	16.4	23.3	
Fallow bamboo	Total (MC + EBT discrepancy)	49.4	49.0	34.8	64.2	
Forest bamboo	Total (MC + EBT discrepancy)	25.1	25.1	21.7	28.6	

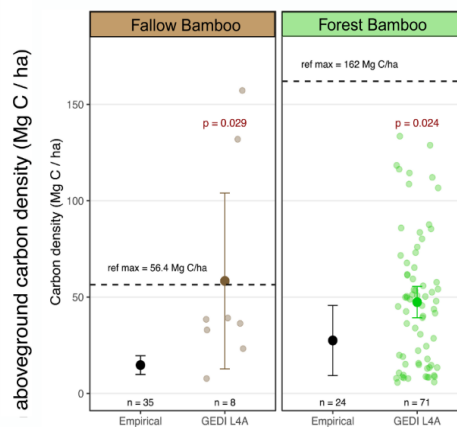
439 **Table 2. Key models and uncertainty components used to extrapolate bamboo carbon bias**
 440 **across Xishuangbanna.** (A) Class-specific RH50–RH98 regressions fitted to GEDI Level 2
 441 footprints (bamboo vs. tree) and used to predict RH50 from a 10-m canopy height layer (treated
 442 operationally as RH98) for regional application of the EBT biomass model. (B) Systematic
 443 discrepancy between biomass/carbon estimated from the simplified RH-only EBT formulation
 444 (Eq. 1) and the published GEDI Level 4A carbon at footprints where Level 2 and Level 4

445 observations overlap, summarized as residual carbon ($C_{est,i} - C_{L4A,i}$). (C) Misclassification-
446 aware GEDI estimates of bamboo carbon bias relative to empirical benchmarks, obtained by
447 posterior Monte Carlo resampling that propagates uncertainty in whether mapped-bamboo
448 footprints are truly bamboo (conditional on the mapped label) and recomputes GEDI–empirical
449 mean differences for forest- and fallow-bamboo contexts. The “Total” component combines the
450 misclassification-aware GEDI–empirical bias with the EBT-formulation discrepancy in Panel B
451 to produce the bias estimates used for regional mapping. Panel C reports bias estimates
452 conditional on footprints mapped as bamboo (false positives are removed); omission errors (false
453 negatives) are not propagated (see Discussion).

a Bamboo in Forest vs Fallow (Predicted)



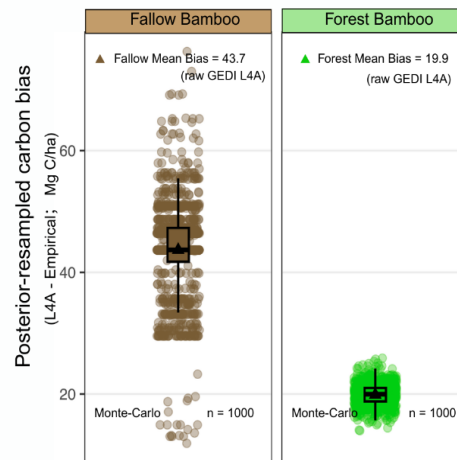
Carbon Overestimation



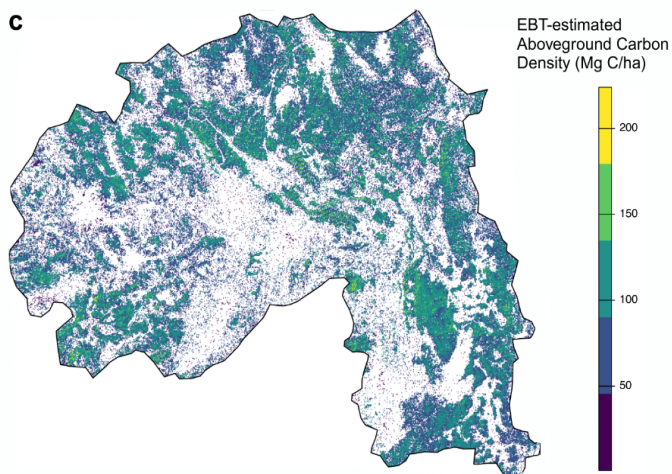
b Mean Carbon Bias as % GEDI-Estimated Value



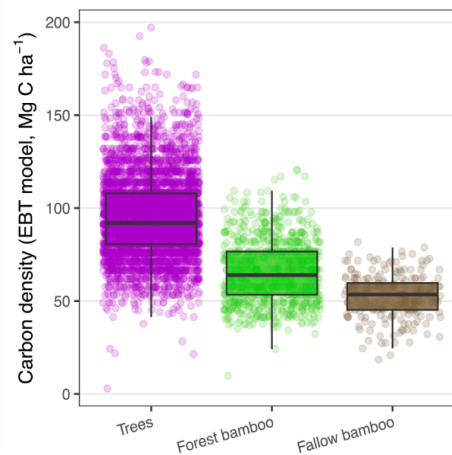
Misclassification-aware Carbon Bias



c



EBT-based carbon estimates



455 **Fig 4. Quantifying carbon overestimation when structurally distinct evergreen broadleaf**
456 **forests are treated with GEDI's single PFT-based EBT biomass model in Xishuangbanna.**

457 **(a)** Predicted bamboo pixels were separated into forest versus fallow contexts using a 3×3
458 neighborhood rule (forest bamboo: ≥ 5 bamboo/tree neighbors; fallow bamboo: < 5). GEDI L4A
459 aboveground carbon density for each bamboo type is compared with regional empirical reference
460 values from Yuen et al. (2017). Error bars show t-based 95% confidence intervals for the group
461 mean carbon density ($\text{mean} \pm t \cdot \text{SD}/\sqrt{n}$); p-values are from one-sided Welch's t-tests (H_1 : GEDI >
462 empirical). A re-derivation of the literature summary statistics is provided in Appendix D.

463 **(b)** Uncertainty arising from bamboo misclassification was propagated using Monte Carlo
464 resampling, in which footprint-level "true" land-cover classes were repeatedly drawn from
465 posterior probabilities derived from the random-forest classifier. These misclassification-aware
466 bias distributions were combined with bias arising from GEDI's simplified EBT biomass
467 formulation based on RH98 and RH50 (Eq.1 applied in GEDI L4A for Xishuangbanna; Kellner
468 et al., 2023) to estimate mean percent carbon overestimation across the region. Fallow bamboo
469 exhibits larger absolute carbon bias. **(c)** Region-wide EBT-based aboveground carbon estimates
470 derived from the simplified EBT RH model. The boxplots summarize the resulting EBT-based
471 carbon densities for pixels classified as trees, forest bamboo, and fallow bamboo. Carbon
472 uncertainties are summarized in Table 2.

473 Comparisons between GEDI Level 4A estimates and empirical bamboo benchmarks
474 reveal systematic overestimation of aboveground carbon density in both bamboo contexts (Fig.
475 4a). GEDI-derived carbon densities for fallow bamboo were significantly higher than empirical
476 values (one-sided Welch's *t*-test, H1: GEDI > empirical, $p = 0.029$), and forest bamboo exhibited
477 an even larger positive deviation from its corresponding benchmark ($p = 0.024$). To limit the
478 disproportionate influence of extreme GEDI footprint values on summary statistics, we excluded
479 outliers in GEDI L4A bamboo carbon density (forest vs fallow) using the Tukey rule (values
480 outside $1.5 \times \text{IQR}$) prior to visualization and statistical comparison.

481 The magnitude of overestimation also differed between bamboo contexts: fallow bamboo
482 showed higher absolute carbon bias (43.7 Mg ha^{-1} versus 19.9 Mg ha^{-1}) than forest bamboo with
483 greater carbon uncertainty, consistent with its lower canopy height and higher landscape
484 complexity observed by GEDI (Fig. 3). In contrast, tree-dominated forests exhibited substantially
485 higher carbon densities overall, as expected for woody evergreen broadleaf systems. Together,
486 these results indicate that applying a single EBT model to bamboo-dominated forests inflates
487 carbon estimates and obscures meaningful variation between bamboo forest and fallow systems.

488 Accounting for land-cover misclassification using posterior Monte Carlo resampling did
489 not eliminate the inferred carbon bias in bamboo-dominated systems (Fig. 4b). Across 1,000
490 posterior resampling iterations, mean aboveground carbon bias remained consistently positive
491 for both bamboo contexts. Fallow bamboo exhibited a mean overestimation of $44.1 \text{ Mg C ha}^{-1}$,
492 whereas forest bamboo showed a larger mean bias of $19.8 \text{ Mg C ha}^{-1}$, closely matching the naïve
493 footprint-level estimates (triangles representing mean values). The narrow spread of the Monte
494 Carlo distributions for forest bamboo indicates that plausible misclassification errors are
495 insufficient to explain the observed magnitude of overestimation, while fallow bamboo with

496 wider distributions suggests higher uncertainty in AGBD estimates likely due to mixed
497 influences from rubber plantations and agroforestry mosaics. These results confirm that the
498 positive carbon bias arises from structural mismatch between bamboo and the EBT biomass
499 model rather than from classification uncertainty alone.

500 Having established that bamboo-specific carbon overestimation persists after explicitly
501 accounting for classification uncertainty at the footprint level, we next examined how this bias
502 propagates spatially across Xishuangbanna. To support regional extrapolation, we quantified
503 class-specific relationships between RH50 and RH98 from GEDI Level 2 observations and
504 evaluated the discrepancy between AGBD predicted by a simplified, RH-only EBT formulation
505 and published GEDI Level 4A estimates. These components were combined to propagate
506 uncertainty into regional estimates of absolute and percent carbon bias. Inputs to this propagation
507 include (1) class-specific RH50–RH98 regressions, (2) the EBT–L4A biomass discrepancy, and
508 (3) land-cover misclassification adjustments; the height dependence of the EBT structural
509 discrepancy is diagnosed in Appendix E. Table 2 summarizes the regression equations and the
510 full bias decomposition.

511 Mapping the percentage carbon bias revealed pronounced spatial heterogeneity across
512 bamboo-dominated landscapes (Fig. 4b). Relative overestimation frequently exceeded 40% in
513 bamboo-rich areas, with systematic differences between fallow and forest bamboo (Fig. 4c).
514 Because percentage bias is defined relative to EBT-estimated carbon density (Eq. 3), spatial
515 variation reflects the joint effects of context-specific absolute bias (ΔC_i) and variation in the
516 underlying EBT-estimated carbon baseline ($AGC_{EBT,i}$). Overall, these spatial patterns reinforce
517 the footprint-level inference that GEDI’s EBT model systematically overestimates bamboo

518 aboveground carbon density, while the magnitude of overestimation varies predictably with
519 bamboo stand context.

520 **4. Discussion**

521 **4.1 Grassy trees drive carbon bias in tree-calibrated biomass models**

522 Our results demonstrate systematic biomass and carbon bias in GEDI Level 4A products over
523 evergreen broadleaf forests in Xishuangbanna, tropical China. This bias persists after Bayesian
524 correction for bamboo–tree classification uncertainty, indicating that it arises primarily from
525 structural mismatch between bamboo-dominated canopies and GEDI’s global EBT biomass
526 model, rather than from misclassification alone. Understanding this bias therefore requires
527 examining how the operational EBT model was trained and applied.

528 The GEDI EBT biomass model is trained on globally uneven data, with continental and
529 Southeast Asia among the most underrepresented regions (Kellner et al. 2023). Although
530 calibration relies on simulated GEDI waveforms linked to evergreen broadleaf forest field plots
531 from 21 countries, only 326 simulated waveforms support the EBT stratum in South Asia,
532 compared with 3,441 in South America and 834 in Africa (Kellner et al. 2023, Table 1). As a
533 result, statistical support for EBT biomass estimation in tropical Asia is substantially weaker than
534 in other regions, a limitation particularly relevant for Xishuangbanna at the northern margin of
535 Southeast Asia.

536 Consistent with this limitation, the CEOS Land Product Validation framework recognizes
537 that, where training data are sparse, biomass estimates must rely on models trained in other
538 regions or plant functional types, implicitly assuming geographic and structural transferability
539 (Duncanson et al. 2022; Kay 2021). While necessary for global coverage, this assumption
540 introduces systematic bias when canopy structure and species composition deviate from those

541 represented in training data, such as bamboo-dominated secondary forests relative to tree-
542 dominated evergreen broadleaf forests.

543 Ecological studies consistently show that bamboo dominance produces pronounced
544 departures from woody evergreen broadleaf forest structure, including reduced tree basal area,
545 altered gap dynamics, suppressed or redirected tree regeneration, and demographic processes
546 driven by clonal expansion and episodic flowering–dieback cycles (Tabarelli and Mantovani
547 2000; Lima et al. 2012; Fadrique et al. 2021; Zheng and Lv 2023). These stands therefore
548 represent fundamentally different canopy architectures, consistent with the structural contrasts
549 observed in GEDI footprint-level LiDAR data (Fig. 3).

550 Mechanistically, bamboo differs from trees in ways that directly violate woody biomass
551 allometries. Bamboo lacks secondary growth, forms hollow culms, and expands clonally through
552 rhizomes (Makita 1998). Whereas tree biomass accumulates through continued radial growth of
553 woody stems—supporting power-law relationships between DBH, height, and biomass (Enquist
554 et al. 1999; Chave et al. 2014)—individual bamboo culms reach near-final height and diameter
555 within a single growing season. Consequently, stand-level biomass increases primarily through
556 culm recruitment and turnover rather than through growth of existing individuals (Yen 2016).
557 Although culm-level DBH–height allometries exist (Sharma et al. 2025), estimating bamboo
558 stand biomass requires explicit consideration of colony structure and culm age composition,
559 complicating direct scaling from tree-based models (Zhang et al. 2014).

560 These biological differences intersect directly with GEDI’s biomass retrieval logic. GEDI
561 Level 4A relies heavily on relative height (RH) metrics (e.g., RH50, RH98), implicitly assuming
562 that waveform-derived vertical structure provides a consistent proxy for woody biomass across
563 vegetation types (Kellner et al. 2023). In bamboo-dominated canopies, however, RH metrics may

564 reflect dense foliage and hollow culm packing rather than woody stem mass, yielding an RH–
565 AGBD relationship that deviates from the tree-calibrated EBT relationship assumed by GEDI
566 L4A. Our height-stratified diagnostics (Fig. E.1; Method E) confirm that residual bias varies
567 systematically with canopy height, indicating structural bias intrinsic to applying the EBT
568 formulation to bamboo-dominated canopies.

569 Taken together, these considerations indicate that systematic bias is expected when
570 GEDI’s global, RH-based EBT biomass model is applied to bamboo-dominated forests. Our
571 analyses in Fig. 3 and Fig. 4 were designed to diagnose and quantify this bias, rather than to
572 challenge the validity of GEDI products per se. Instead, our results align with the CEOS
573 validation framework by emphasizing transparency in model assumptions and responsible
574 interpretation of satellite-derived biomass estimates (Kay 2021). In bamboo-dominated forests,
575 botanical traits such as clonal growth, hollow culms, and rapid culm maturation fundamentally
576 shape canopy structure and biomass accumulation and must be considered when interpreting RH-
577 based biomass products.

578 **4.2 Scope, assumptions, and diagnostic interpretation**

579 Diagnostic scope of the case study

580 This analysis is based on a single regional case study and is designed to develop a workflow for
581 diagnosing and quantifying systematic bias in GEDI biomass estimates, rather than to provide
582 definitive estimates of absolute biomass or carbon density differences between bamboo- and
583 tree-dominated forests. Accordingly, the magnitude of bias we report should be interpreted as an
584 order-of-magnitude reference rather than a universal correction factor. Our results are derived
585 from one ecosystem and one broad contrast (bamboo versus evergreen broadleaf trees), and the
586 magnitude of bias is therefore expected to vary across regions, species compositions, and

587 successional contexts. Classification relies on single-year Sentinel-2 composites because of
588 limited GEDI coverage in Xishuangbanna; incorporating time-series features would likely
589 improve robustness by leveraging temporal patterns of phenological signals.

590 Benchmarking strategy and data constraints

591 A second limitation is the absence of site-specific ground-based biomass measurements in
592 Xishuangbanna. While direct field measurements would provide the strongest benchmark,
593 existing literature values are themselves uncertain. Our analysis therefore does not aim to
594 establish a precise “true” biomass baseline. Instead, our conclusions rest on internally consistent
595 comparisons among GEDI-derived structural metrics, GEDI’s operational EBT biomass model,
596 and published empirical ranges (Yuen et al. 2017). The convergence of structural, ecological,
597 and modeling evidence supports our central conclusion that applying tree-centric, RH-based
598 biomass models to bamboo-dominated canopies introduces systematic bias.

599 Modeling assumptions and uncertainty treatment

600 Our uncertainty propagation relied on standard but simplifying assumptions. We treated the
601 visually interpreted confusion matrix as an observation model and assumed that misclassification
602 rates were stationary and that errors were conditionally independent given the mapped label. We
603 used the same posterior resampling framework throughout the analysis: for footprint-level
604 structural contrasts (Fig. 3; Table 1), we propagated uncertainty across all mapped classes to
605 assess how tree–bamboo differences changed under plausible label reassignment; for the
606 empirical-benchmark carbon comparison (Fig. 4; Table 2), we propagated uncertainty
607 conditional on the mapped bamboo subset to maintain alignment with bamboo-specific
608 benchmarks and context stratification (forest vs fallow). A fully landscape-level correction that
609 also recovered omission errors (false negatives) would require reconstructing latent true-class

610 composition across all land-cover classes and re-deriving bamboo context within each
611 resampling iteration, which was beyond the scope of this workflow-focused case study.

612 Taken together, these limitations do not weaken our conclusions, because our objective is
613 diagnostic rather than predictive: to identify the presence, direction, and structural origin of
614 systematic bias in GEDI biomass estimates. Importantly, our contribution extends beyond bias
615 identification. Using bamboo as a test case, we show that explicitly stratifying structurally
616 distinct canopies using texture and structural features transforms the same remote-sensing data
617 from a source of bias into an opportunity for refinement. This logic generalizes beyond bamboo
618 to other vegetation types whose structure departs from global-model assumptions, including
619 other grassy trees such as palms and bananas, and to fragmented agroforestry systems. We next
620 outline applications that demonstrate how this workflow supports more responsible, bias-aware
621 use of GEDI and related LiDAR products.

622 **4.3 From regional case study to scalable global applications**

623 Our workflow highlights the complementarity between horizontal information from optical
624 imagery (spatial distribution and land-use context) and vertical information from LiDAR
625 (canopy structure), both of which are necessary to interpret ecosystem function and avoid
626 misrepresenting structurally distinct canopy types in biomass estimation. More broadly,
627 stratification is a problem of scale selection under a coverage–accuracy trade-off: the features
628 that best separate target vegetation classes often emerge at intermediate spatial or temporal
629 scales, where signals can be detected consistently across heterogeneous landscapes. These results
630 further emphasize that remote-sensing–derived products must be interpreted with biological
631 awareness, because trait-driven deviations that are negligible at global scales can dominate local
632 error budgets where those traits are prevalent.

633 One immediate application of this framework is place-based monitoring in structurally
634 complex landscapes. By combining Sentinel-2 optical imagery with GEDI LiDAR structure, the
635 workflow enables explicit identification and mapping of locally relevant canopy strata or species
636 of interest, such as grassy trees (e.g., bamboo, palms, and bananas) that are fast-growing yet
637 frequently conflated with trees in global products (Zheng and Lu 2025). This capability is
638 particularly important in fragmented forests and agroforestry mosaics, where compositional
639 changes occur over spatial scales too fine for coarse plant functional type classifications. Rather
640 than relying solely on globally defined PFTs, the workflow allows locally meaningful strata to be
641 detected using region-specific training data and local knowledge, supporting community-led
642 monitoring in heterogeneous landscapes.

643 A second application lies in improving measurement, reporting, and verification (MRV)
644 of forest carbon. Once structurally distinct vegetation types are explicitly identified, the same
645 framework enables more targeted carbon accounting by reducing systematic bias that arises
646 when biomass models are applied beyond their domain of transferability. This is especially
647 relevant for REDD+ and other carbon-crediting or monitoring programs operating in secondary
648 forests, forest–agriculture transition zones, and agroforestry systems, where grassy trees are often
649 promoted or maintained through human management and can contribute substantially to
650 landscape-level biomass yet remain poorly represented in global allometric frameworks (Tang et
651 al. 2025). By separating bias driven by structural mismatch from true spatial variation in
652 biomass, the workflow supports more transparent MRV.

653 Beyond carbon accounting, stratified structure–biomass estimates can guide targeted
654 regeneration, restoration, and management interventions. In multilayered systems such as cocoa-
655 and coffee-based agroforestry, LiDAR-enabled separation of canopy layers—combined with

656 spectral or phenological indicators—can help identify vegetation components that contribute
657 most to productivity, microclimate buffering, or climate resilience (Atalaya-Marin et al. 2025;
658 Pippi et al. 2025). More broadly, our stratify–quantify–scale approach can be extended to track
659 successional trajectories by separating shifts in canopy species composition from disturbance-
660 driven changes in canopy structure, enabling interventions tailored to different successional
661 stages.

662 **Conclusion**

663 As remote sensing technologies advance and global products like GEDI become more accessible,
664 they create new opportunities while also increasing the risk of misinterpretation. Responsible use
665 therefore requires bias-aware interpretation grounded in ecology, organismal biology, land-use
666 history, and local knowledge. Using bamboo as an example, we demonstrate that biologically
667 informed canopy stratification can diagnose and quantify systematic bias in GEDI’s tree-
668 calibrated biomass estimates and provide a practical workflow for doing so. Looking ahead,
669 conceptual and technical advances can further improve the decision value of global products by
670 specifying when stratification is needed, which signals reliably distinguish canopy types, and
671 what validation is sufficient for targeted applications.

672 **Data Availability Statement**

673 All remote-sensing datasets used in this study are publicly available. Data sources and processing
674 details, together with all analysis scripts, are provided at
675 https://github.com/aiyuz/GEDI_Carbon_Workflow.

676 **Author Contributions**

677 AZ, YY, and ML jointly developed the conceptual framework of carbon discrepancy from
678 structural heterogeneity. AZ implemented the workflow, analyzed the data, and prepared all

679 figures and tables. AZ wrote the manuscript and compiled the scripts, with revisions from ML
680 and YY.

681 **Declaration of interests**

682 The authors declare no competing interests.

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688 **Declaration of generative AI and AI-assisted technologies**

689 During the preparation of this work, the authors used ChatGPT (version 5.2) to help organize and
690 annotate scripts for publication and to check the language and grammar of the manuscript. After
691 using this service, the authors reviewed and edited the content and take full responsibility for the
692 content of the published article.

693 References

- 694 Atalaya-Marin, Nilton, Teiser Sanchez-Fuentes, Malluri Goñas, et al. 2025. “Estimation of
695 Aboveground Biomass and Carbon Sequestration in a Cocoa Agroforestry System Using
696 UAV-LiDAR in Northwestern Peru.” *Remote Sensing Applications Society and
697 Environment* 40 (101750): 101750.
- 698 Avalos, Gerardo, Milena Cambronero, and Carolina Alvarez-Vergnani. 2022. “Allometric
699 Models to Estimate Carbon Content in Arecaceae Based on Seven Species of Neotropical
700 Palms.” *Frontiers in Forests and Global Change* 5 (July).
701 <https://doi.org/10.3389/ffgc.2022.867912>.
- 702 Bruening, Jamis, Paul May, John Armston, and Ralph Dubayah. 2023. “Precise and Unbiased
703 Biomass Estimation from GEDI Data and the US Forest Inventory.” *Frontiers in Forests
704 and Global Change* 6 (1149153). <https://doi.org/10.3389/ffgc.2023.1149153>.
- 705 Camargo García, Juan Carlos, Angela Maria Arango Arango, and Long Trinh. 2023. “The
706 Potential of Bamboo Forests as a Carbon Sink and Allometric Equations for Estimating
707 Their Aboveground Biomass.” *Environment Development and Sustainability* 26 (8): 20159–
708 20187.
- 709 Carvalho, Anelena L. de, Bruce W. Nelson, Milton C. Bianchini, Daniela Plagnol, Tatiana M.
710 Kuplich, and Douglas C. Daly. 2013. “Bamboo-Dominated Forests of the Southwest
711 Amazon: Detection, Spatial Extent, Life Cycle Length and Flowering Waves.” *PloS One* 8
712 (1): e54852.
- 713 Chave, Jérôme, Maxime Réjou-Méchain, Alberto Búrquez, et al. 2014. “Improved Allometric
714 Models to Estimate the Aboveground Biomass of Tropical Trees.” *Global Change Biology*
715 20 (10): 3177–3190.
- 716 Descals, Adrià. 2023. “High-Resolution Global Map of Closed-Canopy Coconut Palm.” *Earth
717 System Science Data* 15: 3991–4010.
- 718 Drusch, M., U. Del Bello, S. Carlier, et al. 2012. “Sentinel-2: ESA’s Optical High-Resolution
719 Mission for GMES Operational Services.” *Remote Sensing of Environment* 120 (May): 25–
720 36.
- 721 Dubayah, Ralph, John Armston, Sean P. Healey, et al. 2022. “GEDI Launches a New Era of
722 Biomass Inference from Space.” *Environmental Research Letters* 17 (9): 095001.
- 723 Duncanson, Laura, James R. Kellner, John Armston, et al. 2022. “Aboveground Biomass
724 Density Models for NASA’s Global Ecosystem Dynamics Investigation (GEDI) Lidar
725 Mission.” *Remote Sensing of Environment* 270 (112845): 112845.
- 726 Enquist, Brian J., Geoffrey B. West, Eric L. Charnov, and James H. Brown. 1999. “Allometric
727 Scaling of Production and Life-History Variation in Vascular Plants.” *Nature* 401 (6756):
728 907–911.

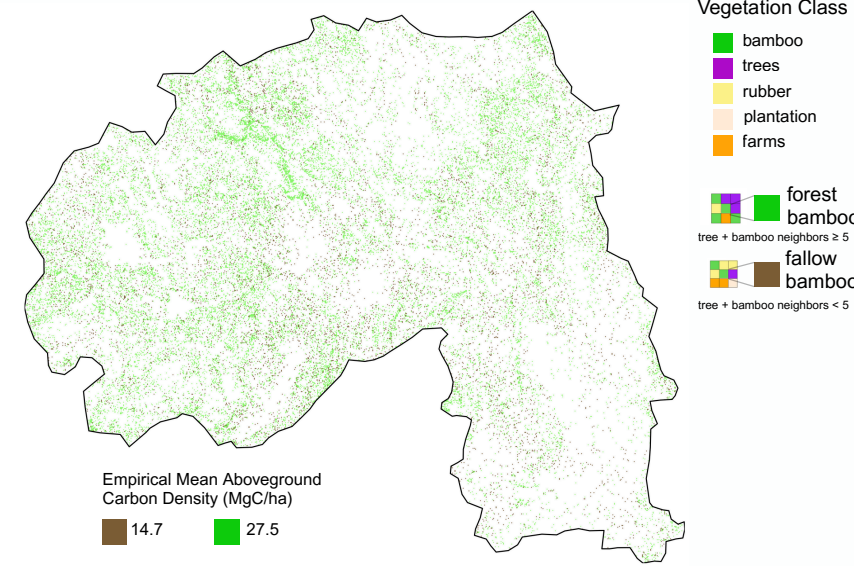
- 729 Fadrique, Belen, Daniel Gann, Bruce W. Nelson, Sassan Saatchi, and Kenneth J. Feeley. 2021.
730 “Bamboo Phenology and Life Cycle Drive Seasonal and Long-term Functioning of
731 Amazonian Bamboo-dominated Forests.” *The Journal of Ecology* 109 (2): 860–876.
- 732 Fadrique, Belén, Joseph W. Veldman, James W. Dalling, et al. 2020. “Guidelines for Including
733 Bamboos in Tropical Ecosystem Monitoring.” *Biotropica* 52 (3): 427–443.
- 734 Heinrich, Viola H. A., Ricardo Dalagnol, Henrique L. G. Cassol, et al. 2021. “Large Carbon Sink
735 Potential of Secondary Forests in the Brazilian Amazon to Mitigate Climate Change.”
736 *Nature Communications* 12 (1): 1785.
- 737 Huang, Xuying, Weimin Ju, Zhanghua Xu, and Jing Li. 2024. “A Novel Method for Mapping
738 Moso Bamboo Forests Using Remote Sensing Data with the Consideration of Phenological
739 Status.” *IEEE Transactions on Geoscience and Remote Sensing: A Publication of the IEEE
740 Geoscience and Remote Sensing Society* 62: 1–18.
- 741 Humeau-Heurtier, Anne. 2019. “Texture Feature Extraction Methods: A Survey.” *IEEE Access:
742 Practical Innovations, Open Solutions* 7: 8975–9000.
- 743 Huy, Bao, Giang Thi Thanh, Krishna P. Poudel, and Hailemariam Temesgen. 2019. “Individual
744 Plant Allometric Equations for Estimating Aboveground Biomass and Its Components for a
745 Common Bamboo Species (*Bambusa Procer* A. Chev. And A. Camus) in Tropical Forests.”
746 *Forests* 10 (4): 316.
- 747 IPCC. 2006. *Intergovernmental Panel on Climate Change Guidelines for National Greenhouse
748 Gas Inventories. Chapter 4: Forest Lands*. [https://www.ipcc-
749 nggip.iges.or.jp/public/mtdocs/pdfiles/LUCF_EGPM_Meeting_Report_Final.pdf](https://www.ipcc-nggip.iges.or.jp/public/mtdocs/pdfiles/LUCF_EGPM_Meeting_Report_Final.pdf).
- 750 Jia, Xiaowei, Ankush Khandelwal, Kimberly M. Carlson, et al. 2020. “Automated Plantation
751 Mapping in Southeast Asia Using MODIS Data and Imperfect Visual Annotations.” *Remote
752 Sensing* 12 (4): 636.
- 753 Kay, Heather. 2021. “Aboveground Woody Biomass Product Validation Good Practices
754 Protocol.” Preprint, Land Product Validation Subgroup (Working Group on Calibration and
755 Validation, Committee on Earth Observation Satellites).
756 <https://doi.org/10.5067/DOC/CEOSWGCV/LPV/AGB.001>.
- 757 Kellner, James R., John Armston, and Laura Duncanson. 2023. “Algorithm Theoretical Basis
758 Document for GEDI Footprint Aboveground Biomass Density.” *Earth and Space Science
759 (Hoboken, N.J.)* 10 (4). <https://doi.org/10.1029/2022ea002516>.
- 760 Kotowska, Martyna M. 2015. “Quantifying above- and Belowground Biomass Carbon Loss with
761 Forest Conversion in Tropical Lowlands of S Umatra (I Ndonesia).” *Global Change Biology*
762 21: 3620–3634.
- 763 Lang, Nico, Nikolai Kalischek, John Armston, Konrad Schindler, Ralph Dubayah, and Jan Dirk
764 Wegner. 2022. “Global Canopy Height Regression and Uncertainty Estimation from GEDI
765 LIDAR Waveforms with Deep Ensembles.” *Remote Sensing of Environment* 268 (112760):

- 766 112760.
- 767 Li, Hongmei, Youxin Ma, T. Mitchell Aide, and Wenjun Liu. 2008. “Past, Present and Future
768 Land-Use in Xishuangbanna, China and the Implications for Carbon Dynamics.” *Forest
769 Ecology and Management* 255 (1): 16–24.
- 770 Li, Longwei, Nan Li, Dengsheng Lu, and Yuyun Chen. 2019. “Mapping Moso Bamboo Forest
771 and Its on-Year and off-Year Distribution in a Subtropical Region Using Time-Series
772 Sentinel-2 and Landsat 8 Data.” *Remote Sensing of Environment* 231 (111265): 111265.
- 773 Lima, Renato A. F., Débora C. Rother, Ana E. Muler, Igo F. Lepsch, and Ricardo R. Rodrigues.
774 2012. “Bamboo Overabundance Alters Forest Structure and Dynamics in the Atlantic Forest
775 Hotspot.” *Biological Conservation* 147 (1): 32–39.
- 776 Lin, Di, Mario Elia, Onofrio Cappelluti, et al. 2025. “From Spaceborne LiDAR to Local
777 Calibration: GEDI’s Role in Forest Biomass Estimation.” *Remote Sensing* 17 (16): 2849.
- 778 Makita, Akifumi. 1998. “The Significance of the Mode of Clonal Growth in the Life History of
779 Bamboos.” *Plant Species Biology* 13 (2-3): 85–92.
- 780 Mohammadpour, Pegah, Domingos Xavier Viegas, and Carlos Viegas. 2022. “Vegetation
781 Mapping with Random Forest Using Sentinel 2 and GLCM Texture feature—A Case Study
782 for Lousã Region, Portugal.” *Remote Sensing* 14 (18): 4585.
- 783 Nath, Arun Jyoti, Rattan Lal, and Ashesh Kumar Das. 2015. “Managing Woody Bamboos for
784 Carbon Farming and Carbon Trading.” *Global Ecology and Conservation* 3 (January): 654–
785 663.
- 786 Navarro, M. N. V., C. Jourdan, T. Sileye, et al. 2008. “Fruit Development, Not GPP, Drives
787 Seasonal Variation in NPP in a Tropical Palm Plantation.” *Tree Physiology* 28 (11): 1661–
788 1674.
- 789 Pan, Yude, Richard A. Birdsey, Jingyun Fang, et al. 2011. “A Large and Persistent Carbon Sink
790 in the World’s Forests.” *Science (New York, N.Y.)* 333 (6045): 988–993.
- 791 Pan, Yude, Richard A. Birdsey, Oliver L. Phillips, et al. 2024. “The Enduring World Forest
792 Carbon Sink.” *Nature* 631 (8021): 563–569.
- 793 Pippi, Lorenzo, Michael Alibani, Daniele Antichi, et al. 2025. “Use of Digital Technologies into
794 Agroforestry Systems: A Review.” *Agronomy (Basel, Switzerland)* 15 (12): 2671.
- 795 Qi, Shuhua, Bin Song, Chong Liu, et al. 2022. “Bamboo Forest Mapping in China Using the
796 Dense Landsat 8 Image Archive and Google Earth Engine.” *Remote Sensing* 14 (3): 762.
- 797 Senf, Cornelius, Dirk Pflugmacher, Sebastian Van der Linden, and Patrick Hostert. 2013.
798 “Mapping Rubber Plantations and Natural Forests in Xishuangbanna (southwest China)
799 Using Multi-Spectral Phenological Metrics from MODIS Time Series.” *Remote Sensing* 5
800 (6): 2795–2812.

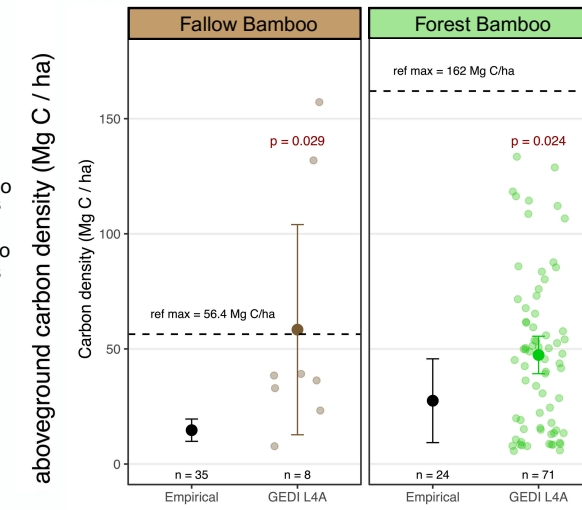
- 801 Sharma, Aastha, Santosh Ayer, Keshav Ayer, et al. 2025. “Empirical Allometric Models for
802 Estimating Aboveground Biomass of *Bambusa Teres* and *Bambusa Tulda* in Non-Forest
803 Areas of Nepal.” *Agroforestry Systems* 99 (5). <https://doi.org/10.1007/s10457-025-01221-9>.
- 804 Song, Xinzhang, Xiaofeng Chen, Guomo Zhou, Hong Jiang, and Changhui Peng. 2017.
805 “Observed High and Persistent Carbon Uptake by Moso Bamboo Forests and Its Response
806 to Environmental Drivers.” *Agricultural and Forest Meteorology* 247 (December): 467–475.
- 807 Tabarelli, Marcelo, and Waldir Mantovani. 2000. “Gap-Phase Regeneration in a Tropical
808 Montane Forest: The Effects of Gap Structure and Bamboo Species.” *Plant Ecology* 148 (2):
809 149–155.
- 810 Tang, Yuzhi, Chao Yang, Haishan Wu, et al. 2025. “Tropical Forest Carbon Offsets Deliver
811 Partial Gains amid Persistent over-Crediting.” *Science (New York, N.Y.)* 390 (6769): 182–
812 187.
- 813 Venkatappa, Manjunatha, Sutee Anantsuksomsri, Jose Alan Castillo, Benjamin Smith, and
814 Nophea Sasaki. 2020. “Mapping the Natural Distribution of Bamboo and Related Carbon
815 Stocks in the Tropics Using Google Earth Engine, Phenological Behavior, Landsat 8, and
816 Sentinel-2.” *Remote Sensing* 12 (18): 3109.
- 817 Wang, Kanglin, Ramanatha Rao, and Hong. 2002. “Diversity and Indigenous Utilization of
818 Bamboo in Xishuangbanna, Yunnan Province, Southwest China.” *Journal of Bamboo and
819 Rattan* 1 (3): 263–273.
- 820 Wang, Yao, Hongliang Fang, Yu Li, Sijia Li, and Hao Tang. 2025. “Validation of the Vertical
821 Plant Area Index Profile Product Derived from GEDI over Global Forest Sites.”
822 *Agricultural and Forest Meteorology* 371 (110612): 110612.
- 823 Xayalath, Singkone, Isao Hirota, Shinsuke Tomita, and Michiko Nakagawa. 2019. “Allometric
824 Equations for Estimating the Aboveground Biomass of Bamboos in Northern Laos.” *Journal
825 of Forest Research* 24 (2): 115–119.
- 826 Yen, Tian-Ming. 2016. “Culm Height Development, Biomass Accumulation and Carbon Storage
827 in an Initial Growth Stage for a Fast-Growing Moso Bamboo (*Phyllostachy Pubescens*).”
828 *Botanical Studies (Taipei, Taiwan)* 57 (1): 10.
- 829 Yang, Qing, Zhu-Biao Duan, Zheng-Liang Wang, Kai-Hong He, Qi-Xiang Sun, and Zhen-Hua
830 Peng. 2008. “Bamboo Resources, Utilization and Ex-Situ Conservation in Xishuangbanna,
831 South-Eastern China.” *Journal of Forestry Research* 19 (1): 79–83.
- 832 Yen, Tian-Ming. 2016. “Culm Height Development, Biomass Accumulation and Carbon Storage
833 in an Initial Growth Stage for a Fast-Growing Moso Bamboo (*Phyllostachy Pubescens*).”
834 *Botanical Studies (Taipei, Taiwan)* 57 (1): 10.
- 835 Yuen, Jia Qi, Tak Fung, and Alan D. Ziegler. 2017. “Carbon Stocks in Bamboo Ecosystems
836 Worldwide: Estimates and Uncertainties.” *Forest Ecology and Management* 393 (June):
837 113–138.

- 838 Yusof, Nurasmalaily, Helmi Zulhaidi Mohd Shafri, and Nur Shafira Nisa Shaharum. 2021. “The
839 Use of Landsat-8 and Sentinel-2 Imageries in Detecting and Mapping Rubber Trees.”
840 *Journal of Rubber Research (Kuala Lumpur, Malaysia)* 24 (1): 121–135.
- 841 Zhang, Houxi, Shunyao Zhuang, Bo Sun, Haibao Ji, Changming Li, and Sai Zhou. 2014.
842 “Estimation of Biomass and Carbon Storage of Moso Bamboo (*Phyllostachys Pubescens*
843 *Mazel Ex Houz.*) in Southern China Using a Diameter–age Bivariate Distribution Model.”
844 *Forestry (London, England)* 87 (5): 674–682.
- 845 Zhang, Jianhou, and Min Cao. 1995. “Tropical Forest Vegetation of Xishuangbanna, SW China
846 and Its Secondary Changes, with Special Reference to Some Problems in Local Nature
847 Conservation.” *Biological Conservation* 73 (3): 229–238.
- 848 Zheng, Aiyu, and Mingzhen Lu. 2025. *Grassy Trees: The Neglected Hybrids for Sustainability*.
849 October 23. [https://www.cell.com/trends/ecology-evolution/abstract/S0169-5347\(25\)00287-](https://www.cell.com/trends/ecology-evolution/abstract/S0169-5347(25)00287-3)
850 [3.](https://www.cell.com/trends/ecology-evolution/abstract/S0169-5347(25)00287-3)
- 851 Zheng, Aiyu, and Jianhua Lv. 2023. “Spatial Patterns of Bamboo Expansion across Scales: How
852 Does Moso Bamboo Interact with Competing Trees?” *Landscape Ecology* 38 (12): 3925–
853 3943.
- 854 Zhu, Hua, Min Cao, and Huabin Hu. 2006. “Geological History, Flora, and Vegetation of
855 Xishuangbanna, Southern Yunnan, China¹.” *Biotropica* 38 (3): 310–317.
- 856

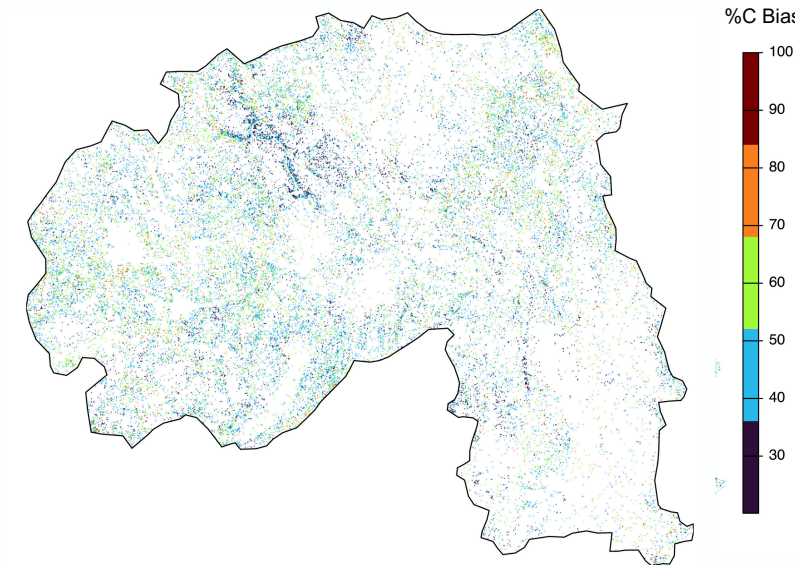
a Bamboo in Forest vs Fallow (Predicted)



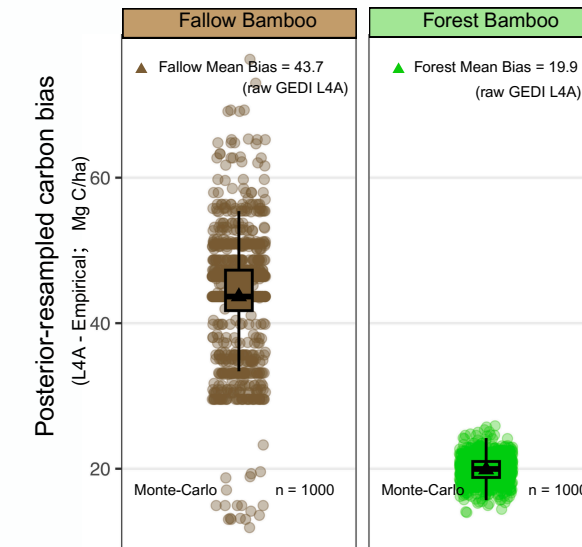
Carbon Overestimation



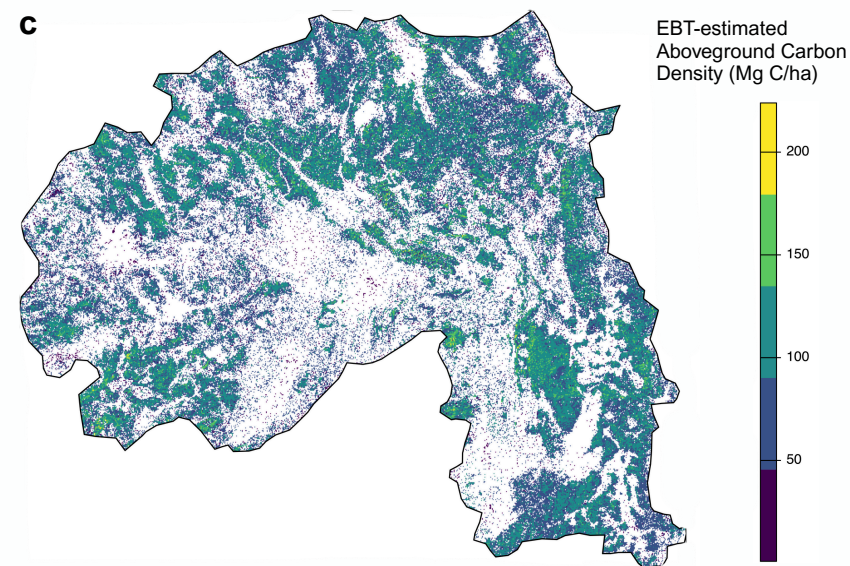
b Mean Carbon Bias as % GEDI-Estimated Value



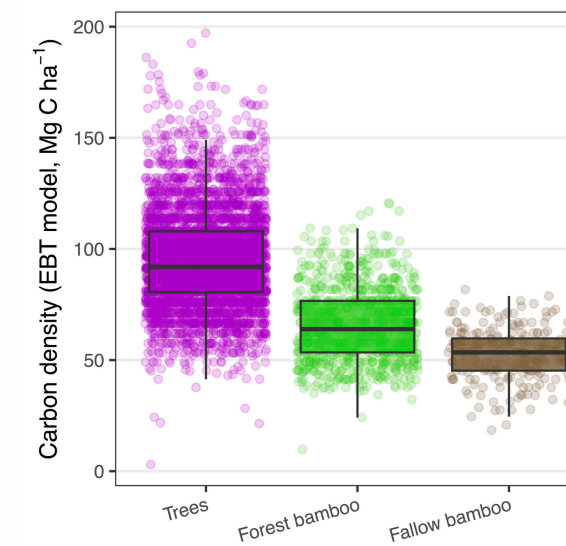
Misclassification-aware Carbon Bias

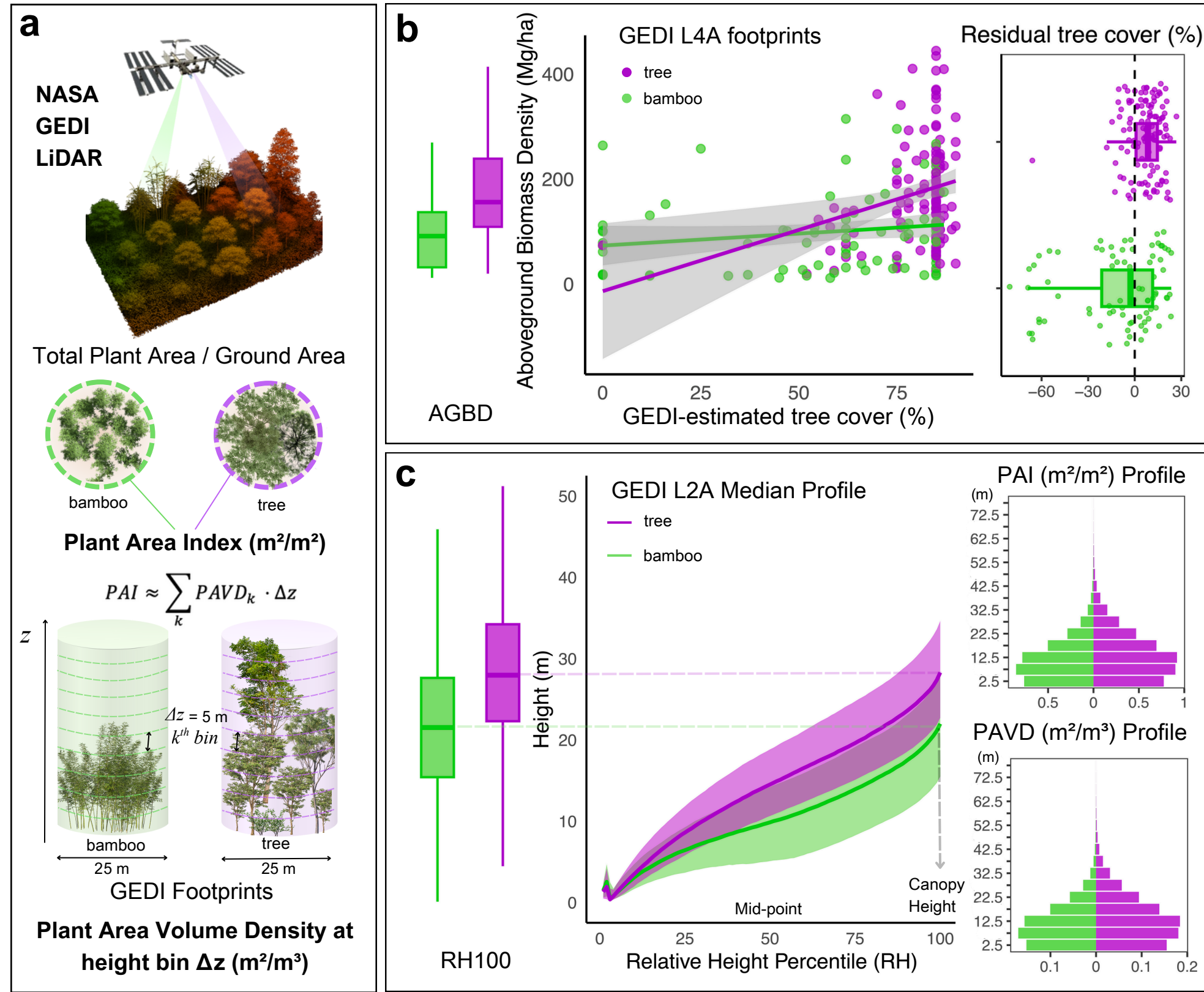


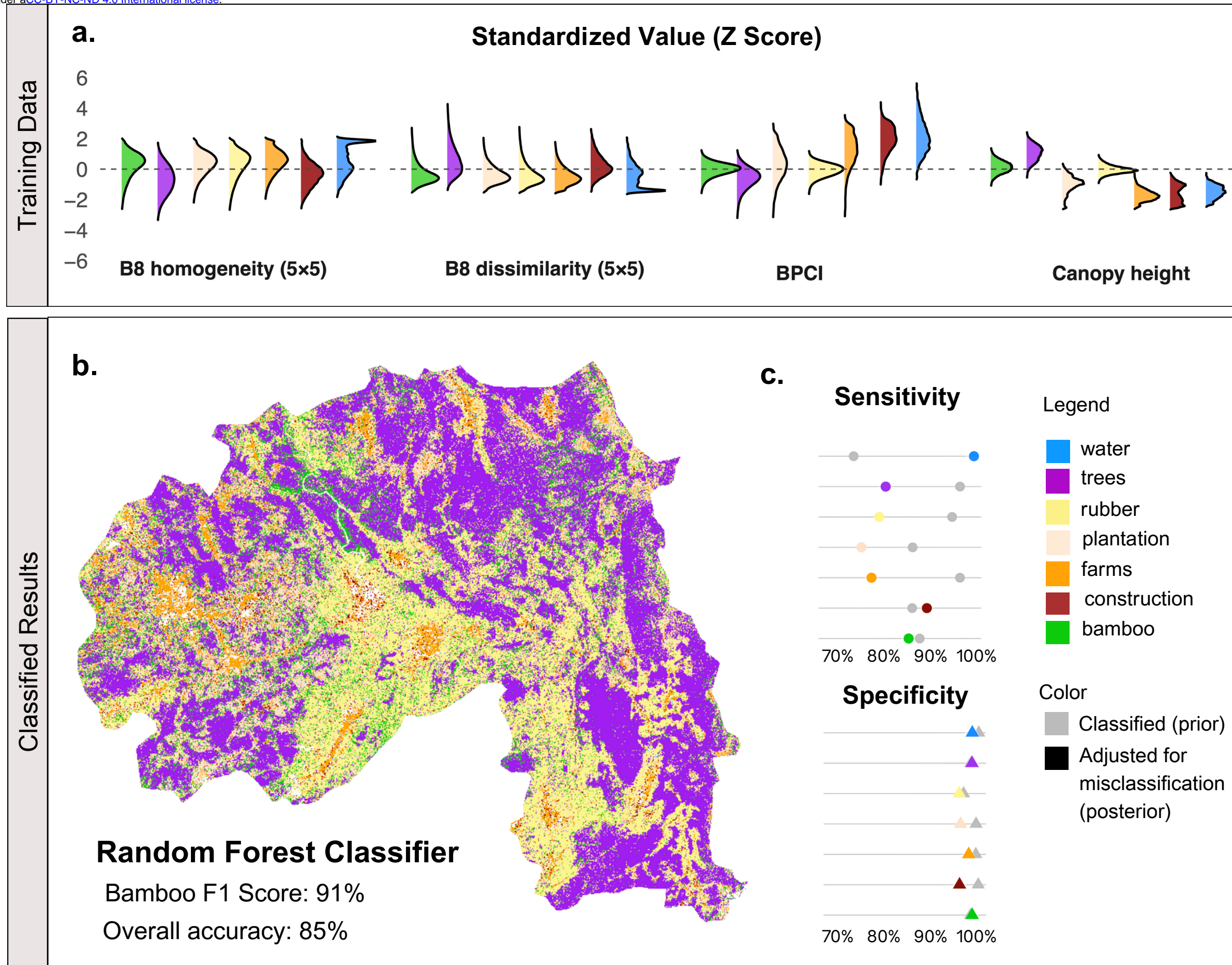
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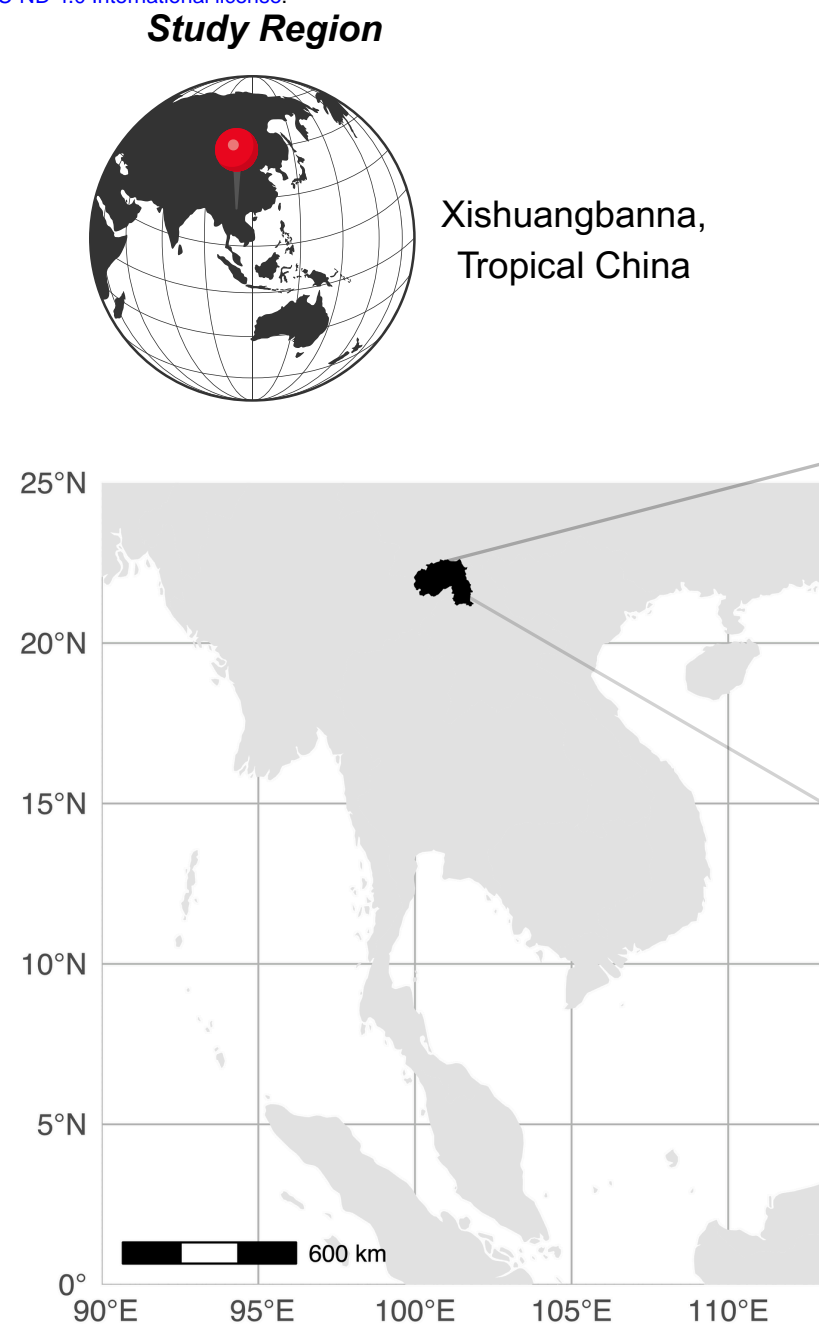


EBT-based carbon estimates





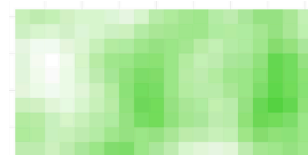




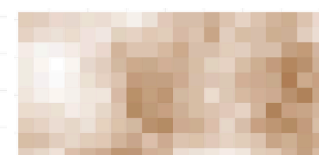
Study Aim

Quantify carbon bias arising from applying a single PFT biomass model to structurally heterogeneous vegetation

Bamboo Map



Δ Carbon (MgC/ha)



GEDI's Source of carbon bias

Bamboo-dominant

tree-dominant



both are labelled
Evergreen Broadleaf Trees



despite significant
Structural Differences

Workflow

Objective 1: Stratify

Classify Bamboo- and Non-bamboo Evergreen Broadleaf Forests using Sentinel-2

Objective 2: Quantify

Quantify Structural Differences under Coarse PFT labeling using GEDI LiDAR

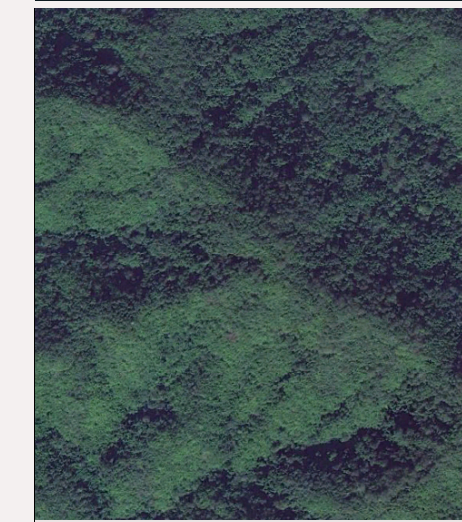
Objective 3: Scaling

Carbon Discrepancy and Other Implications

Training Data

Maxar Optical (<1m) + DEM/DSM

Observation data (Literature)



1.1 Expert Labeling

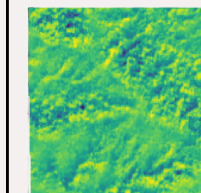
Mapping Data

Sentinel 2 Optical (10 m)

Global Canopy Height (10 m)

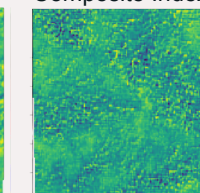
Texture Metrics (GLCM)

NIR



NIR Dissimilarity

Bamboo Probability Composite Index



NIR Homogeneity

1.2 Feature Extraction

Classification Algorithm

1.3 Supervised Classification

1.4 Evaluation and Validation